

RECORD OF DECISON

For

AREA 1 UPLAND AND IN-WATER AREAS FORMER POPE AND TALBOT WOOD-TREATING SITE ST. HELENS, OREGON

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State of Oregon
DEO Department of Environmental Quality

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1. INTRODUCTION

1.1 INTRODUCTION

This Record of Decision (ROD) presents the Oregon Department of Environmental Quality's selected cleanup remedy (remedial action) for Area 1 Upland and contaminated sediments located in Upper Milton Creek and Scappoose Bay adjacent to the former Pope and Talbot wood-treating facility located at 1550 Railroad Avenue in St. Helens, Oregon (Site) (Figure 1). The Site has been divided into several areas. DEQ issued a no further action determination for Area 2 Upland (tax lot 302) on April 16, 2012. This ROD is applicable to the Area 1 Upland (tax lot 300) and four priority sediment areas (Area 1 Dock, Area 2 Dock, Cove Area, and Upper Milton Creek) which are defined in Section 2.1.1 and shown on Figure 2. The selected remedial action includes riverbanks located within the priority sediment areas. The selected remedial action was developed in accordance with Oregon Revised Statutes (ORS) 465.200 *et seq*. and Oregon Administrative Rules (OAR) Chapter 340, Division 122, Sections 0010 through 0115, and is based on the administrative record for this Site.

This ROD summarizes information contained in the DEQ Staff Report (DEQ, 2023), as well as the Remedial Investigation, Updated Supplemental Risk Assessment, Human Health Risk Assessment, Ecological Risk Assessment, and Remedial Investigation/Feasibility Study reports completed under DEQ Order of Consent Number (No.) WMCSR-NWR-95-05 signed April 13, 1995, by Port of St. Helens (now Port of Columbia County) and DEQ. DEQ's environmental cleanup site information (ECSI) system designates the Site as No. 0959. A copy of the Administrative Record Index is attached as Section 13.

In selecting the remedial action presented in this ROD, DEQ considered public input following a comment period on the Staff Report.

1.2 SCOPE AND ROLE OF THE SELECTED REMEDIAL ACTION

The selected remedial action addresses both upland and in-water contamination. In the upland, the selected remedial action addresses the presence of creosote non-aqueous liquid (NAPL) and dissolved phase chemical of concerns (COCs) in groundwater and soil in the Area 1 Upland priority action area (PAA). In the in-water areas, the selected remedial action addresses sediment, porewater and surface water impacted by polycyclic aromatic hydrocarbons (PAHs), metals, creosote sheen, and contaminated woody debris in surface and subsurface sediments within the sediment PAAs and adjacent riverbanks where contaminated soils pose a future recontamination risk. The areas with the greatest potential for impacts to human health and the environment are the focus of the feasibility study and a key focus on the selected remedial action. Additional in-water

areas (e.g., with light petroleum sheen) may need further investigation and cleanup in the future (e.g., Lower Milton Creek).

The selected remedial action for Area 1 Upland includes placement of an impervious surface cap over the entire Area 1 Upland PAA and a permeable reactive barrier (PRB) at the top of the riverbank, adjacent to the Cove Area PAA. The selected upland remedial action is intended to address risks associated with direct contact to soil in the upland, limit stormwater infiltration into the NPAL source area in Area 1 Upland, and limit contaminant migration from the upland source area into the in-water areas via groundwater seeps. The long-term effectiveness of the selected remedial action for Area 1 Upland is contingent upon Institutional Controls (ICs) to mitigate risks to potential future industrial, excavation, and construction workers in the upland.

The selected in-water remedial action is intended to address areas with the highest levels of sediment contamination (hot spots, specifically areas with creosote NAPL and heavy to moderate petroleum sheen). Remedial action objectives (RAOs) were developed to focus treatment in areas with the highest contamination levels posing unacceptable risks to humans and ecological receptors. DEQ anticipates further investigations to support remedial design will be performed.

Contamination in upland soil and groundwater pose unacceptable direct contact and vapor inhalation risks to human health (specifically to potential future industrial, excavation, and construction workers). Sediment contamination within the in-water PAAs exceeds acceptable risk levels for both human health (specifically to subsistence fishers) and for aquatic ecological receptors. In addition to risks associated with dissolved phase COCs, the presence of creosote NAPL and moderate to heavy petroleum sheen presents unacceptable risk to human health and aquatic ecological receptors. "Hot spots" (elevated levels of contamination) are present in all of the PAAs.

The following general actions will be components of the selected remedial action:

- Placement of an impervious surface cap over the entire Area 1 Upland PAA and a PRB at the top of the riverbank, adjacent to the Cove Area PAA.
- Establishment of an easement and equitable servitudes (EES) in the upland to maintain the cap in perpetuity, and implementation of ICs to prevent potential risks to future industrial, excavation, and construction workers associated with groundwater and subsurface soil contamination in Area 1 Upland.
- Treatment, containment, or removal of sediment that contains NAPL and/or emanates moderate to heavy petroleum sheen.
- Treatment, containment, or removal of erodible riverbank soils considered to pose a risk of recontamination to sediments.
- Monitored natural recovery (MNR), which consists of the natural burial of surface sediment contamination posing a lower risk through deposition of suspended sediment from Milton Creek, Columbia, and Multnomah Channel watersheds.
- An Oregon Health Authority (OHA) advisory regarding consumption of fish, shellfish, and crayfish in Multnomah Channel and Scappoose Bay to minimize potential risk to human health until contamination levels protective of human health and fish are achieved.
- Long-term monitoring to assess and document progress of active remedial measures and MNR in achieving RAOs and cleanup levels (CULs) long-term and to address any residual risk.
- Periodic land and water use review.

The selected remedial action consists of the following PAA-specific elements:

- In the Area 1 Upland PAA, placement of an impervious surface cap with an engineered stormwater management system over the entire Area 1 Upland to reduce infiltration in combination with an organoclay PRB to intercept and sequester contaminants in the groundwater: preventing recontamination of the in-water remedy.
- Within the Area 1 Dock PAA, timber pilings and impacted surficial woody debris will be removed and an amended isolation cap will be placed across the sediment containing the highest levels of contamination posing unacceptable risks to human health and aquatic ecological receptors.
- In the Area 2 Dock PAA, impacted surficial woody debris will be removed, shallow surface sediment will be removed near the shoreline, and a sand cap will be placed over residual contamination for enhanced natural recovery (ENR).
- In the Cove Area and Upper Milton Creek PAAs, the shallow sediment with the highest concentrations (creosote NAPL and/or moderate to heavy sheen) will be excavated for offsite disposal and an engineered, amended isolation cap will be placed over the sediment within the PAAs along with protective armoring. The riverbank adjacent to these PAAs will be regraded to remove impacted soil and further reduce the potential for recontamination of the in-water remedy.

The selected remedial action will restore the Site to conditions protective of risks to upland Site workers, people who consume fish and shellfish for recreational and subsistence purposes or are directly exposed to contaminated sediment, and aquatic ecological receptors that consume prey from the Site or are directly exposed to contaminated sediment. The selected remedial action is protective of beneficial water uses from releases of petroleum sheen and prevents recontamination of the in-water remedy from the adjacent groundwater seeps and erodible riverbank soils.

2.1 SITE LOCATION AND LANDUSE

The Site consists of approximately 42 acres of industrial zoned land situated within the southern limits of the City of St. Helens. DEQ approved subdividing the upland portion of the Site into two areas (Area 1 and Area 2) in 2008. Area 1 is 25.15 acres and Area 2 is 17.32 acres. The Site also includes adjacent riverbanks and sediment areas. The United States (U.S.) Army Corps of Engineers (USACE) Ordinary High-Water Mark (OHWM) divides the upland and in-water areas. Land above the OHWM is considered upland and the portions of riverbanks and sediments below the OHWM are considered to be in–water. The selected remedial action is specific to Area 1 Upland PAA and four in-water PAAs discussed further in Section 2.1.1. DEQ's ECSI system designates the Site as No. 0959.

The Site lies on the northwestern bank of Scappoose Bay, near the confluence of Scappoose Bay and the Multnomah Channel. The Multnomah Channel is a 21.5-mile distributary, or branch, of the Willamette River. The channel flows to the north adjacent to the west side of Sauvie Island until it meets the Columbia River near St. Helens. Scappoose Bay is a low-energy surface water embayment just west of Multnomah Channel that includes former and current industrial sites. The confluence of the Multnomah Channel and the Columbia River is approximately 1.3 miles downriver from the Site. Milton Creek is a perennial stream located on the Site's western boundary. Milton Creek discharges to Scappoose Bay at the southwest corner of the Site.

2.1.1 Areas of the Site

The upland ground surface across most of Area 1 is generally between 22 to 25 feet North American Vertical Datum of 1988 (NAVD88). The Site's top of bank elevation with Milton Creek and Scappoose Bay are between 19 and 21 feet NAVD88. The OHWM at the Site generally corresponds to a vertical elevation of approximately 14 feet relative to the NAVD88. As previously discussed, riverbanks above the OHWM are included in Area 1 and riverbanks below OHWM are included with adjacent in-water areas. The riverbanks along Scappoose Bay and Milton Creek have steep to vertical banks. Portions of the riverbank, especially along Milton Creek, have potentially erodible soils that pose a recontamination risk to in-water remedies.

2.1.1.1 Area 1 Upland Priority Action Area

Geographically, Area 1 Upland is located approximately one mile east of Highway 30. Area 1 has a street address of 1550 Railroad Avenue in St. Helens, Oregon, and is tax lot 300 located in Section 9, Township 4 North, Range 1 West of the Willamette Baseline and Meridian in Columbia County (Figure 1). The latitude is 45.8413 degrees North, longitude 122.8094 degrees West. Area 1 Upland is accessed by and located at the eastern end of Railroad Avenue. Area 1 Upland covers 25.15 acres of the southwest portion of the Site and is generally flat with a peak elevation of approximately 27 feet above mean sea level. Nearly all of Area 1 Upland is located within the 100-year floodplain. Area 1 Upland is primarily vacant industrial land with a tenant occupying the former office/shop building. The Area 1 Upland boundary is shown on Figure 2.

The former wood-treating plant facilities, above ground storage tanks (ASTs), transfer tables, loading platforms, and underground portion of the creosote pipeline were located within Area 1 Upland (see Figures 3A and 3B). All facilities except for the former office/shop and a storage shed were removed from Area 1 in the early 1960s. Dredge material generated during the deepening of the Columbia River was placed on Area 1 Upland in the early 1970s. The fill material averages 7 to 8 feet thick over the native soil. Creosote NAPL observed in Area 1 Upland riverbanks adjacent to Milton Creek and Scappoose Bay indicates that some lateral movement of NAPL occurs beneath the Site. The progression of Site features from 1919 through 2013 are shown on Figures 4 through 16. A generalized cross-section showing the relationship between the dredge fill material and native soil and basalt bedrock is shown on Figure 17.

2.1.1.2 Area 1 Dock Priority Action Area

The Area 1 Dock PAA consists of an approximately 700-foot section of shoreline adjacent to Area 1 Upland and in-water area located between the former hog fuel loading hopper and downriver terminus of the former transfer table dock (see Figure 2). The wood decking and lateral bracing was removed from both dock structures in 2013. Over 435 the creosote-treated timber pilings remain, and the wood debris is present in the upper 1 to 2 feet of sediment and as deep as 11 feet below the mudline (bml; "mudline" generally refers to the surface water/sediment interface). The source of surface sediment exhibiting moderate to heavy sheen in the Area 1 Dock PAA is creosote-contaminated and/or treated wood debris associated with historical overwater activities, which has been encountered within 1.85 acres of surface sediment underneath and surrounding the former dock structures.

The Area 1 Dock PAA is closer to the main channel of Scappoose Bay and is subjected to stronger river current, as evidenced by the coarser sandy sediments encountered within the intertidal zone. The area's shoreline appears to be relatively stable to eroding. This indicates the environment is generally preventing the deposition of newer clean sediments and the burial of historical contamination.

2.1.1.3 Area 2 Dock Priority Action Area

The Area 2 Dock PAA consists of an approximately 600-foot section of shoreline adjacent to Area 2 Upland and in-water area located near the downriver property boundary and terminus of the former creosote pipeline, AST, and historical off-loading dock (see Figure 2). The wood decking and lateral bracing was removed from both dock structures. Hundreds of closely spaced creosote-treated timber pilings remain. Moderate to heavy petroleum sheen has been observed near timber pilings within the upper 12 inches of sediment. Residual creosote contaminated wood debris has been encountered within 0.3 acres of surface sediment in the former dock structure area.

The main stem of the Multnomah Channel is reportedly 100 to 150 feet offshore with a bottom elevation of approximately -15 feet NAVD88 immediately downstream of the Site. Based on the presence of contamination in surface sediment only (i.e., upper 12 inches) and the hard compact nature of the mudflats with relatively steep slopes towards the channel, the Area 2 Dock PAA appears to be relatively stable to eroding.

2.1.1.4 Cove Priority Action Area

The Cove PAA consists of approximately 300 feet of shoreline adjacent to Area 1 Upland, the island, and around the peninsula (see Figure 2). The Cove PAA is immediately downgradient of the former wood-treating plant operations. The cove is a manmade area created between 1948 and 1953 by extensive dredging and filling of lowlands. The Cove PAA consisted of a partially vegetated slough prior to dredging.

The elevation of the river bottom within the manmade cove and shallow inlet immediately downriver of the peninsula ranges between approximately 0 and 10 feet NAVD88 within approximately 100 to 300 feet of the shoreline before sloping steeply towards the main channel of Scappoose Bay. The main channel of Scappoose Bay adjacent to the Cove PAA widens to approximately 150 feet with bottom elevations between -7 and -5 feet NAVD88.

The Cove PAA is protected from the erosional effects of large storm and tidal surge and is a sediment depositional area. In 1960, wood treating plant operations ceased which has led to the deposition and burial of the old operational sediment surface throughout the Cove PAA. A sediment deposition rate of 0.2 to 1.4 inches per year (in/yr) is implied based on the presence of 1 to 7 feet of relatively clean fine textured sediment above the old contaminated operational sediment surface.

Creosote-impacted groundwater seeps exhibiting moderate to heavy sheen have been intermittently observed along the Cove PAA riverbank, located west of the peninsula. Some sediments within certain areas of the Cove PAA have been observed to exhibit moderate to heavy or slight petroleum sheen. NAPL within the Cove PAA varies vertically from a few inches to 12 feet bml, with much of the historical creosote contamination located 2 to 3 feet below deposited relatively clean sediment.

2.1.1.5 Upper Milton Creek Priority Action Area

Milton Creek discharges to Scappoose Bay at the southwest corner of the Site (see Figure 2). The Upper Milton Creek PAA consists of a 200-foot section of the east bank of Milton Creek where groundwater seeps exhibiting moderate to heavy creosote sheen are present within a layer of sand near the base of the streambank.

The upper portion of Milton Creek adjacent to the Site is characterized by relatively steep embankments (e.g., up to 1.5H:1V with "H" the cutback distance and "V" the depth). The elevation difference between the top and bottom of the streambank is approximately 15 feet along this portion of Milton Creek. Scouring of the creek channel by seasonal runoff and diurnal tidal cycles appears to limit the accumulation of sediment in this portion of the creek. The sediment bed profile generally consists of 0 to 1 foot of silty organic-rich fluff overlying a hardpan layer of clayey silt with gravel and cobbles.

2.1.1.6 Lower Milton Creek Sediment Area

The lower portion of Milton Creek generally includes the confluence of Milton Creek with Scappoose Bay and is characterized by a depositional sediment environment, with 3 to 10 feet of soft silt and varying amounts of sand overlying Columbia River Basalt Bedrock. Because this area

generally has not had observations of creosote NAPL in sediments and is depositional, the area is not considered a PAA. Additional sampling in this and other areas where petroleum sheen has been observed outside of the PAAs may be needed in the future.

2.1.2 Current and Reasonably Anticipated Future Land Uses

The Upland Area is currently zoned for Heavy Industrial (HI) use as part of the Railroad Avenue Industrial Park. The current and reasonably anticipated future uses of the Upland Area remain Heavy Industrial.

Adjacent land uses include railroad tracks and undeveloped land to the north and City-owned industrial development (former Boise White Paper Mill) to the northeast. Northwest of the Site and north of the railroad tracks are several single-family homes. Milton Creek forms the western Site boundary with the former St. Helens Fiberboard Plant and the Columbia County transfer station located on the west side of Milton Creek. Scappoose Bay and the Multnomah Channel of the Willamette River form the southern and eastern Site boundaries.

2.1.3 Adjacent Cleanup Sites

The Site is bordered to the east and northeast by the more than 200-acre former Boise Cascade Mill/Boise White Paper Mill site (ECSI No. 0014) located at 1300 Kaster Road in St. Helens (Figure 18). The former Boise White Paper Mill property is current and former industrial land with various buildings that were previously the location of pulping, bleaching, and milling operations. The former Boise White Paper Mill property issued a Record of Decision in 2023 for the 15-acre In-Water Sediment area. Additional investigation is anticipated to be performed during Remedial Design.

The former St. Helens Fiberboard Plant site (ECSI No.0091) is located at 1645 Railroad Avenue, west of the Site (Figure 18). The inactive former St. Helens Fiberboard Plant site has been used to manufacture a variety of mineral fiber and wood fiber building products. The site is divided into two operable units: Upland area and Lowland/In-water area. Remedial action in the upland area is complete. An interim remedial action measure is underway at the St. Helens Fiberboard Plant site in the Lowland/In-water area.

2.2 PHYSICAL SETTING

2.2.1 Climate

St. Helens has a temperate marine climate characterized by short, dry summers and wet winters. Between [1](#page-14-0)981 and 2010 ,¹ the average annual precipitation in St. Helens was 46.6 inches. Most precipitation falls between November and May, with average monthly totals ranging from 0.72 to 7.22 inches and the highest in December. The mean annual temperature ranges from approximately 37.1 to 67.2 degrees Fahrenheit (°F).

¹ National Oceanic & Atmospheric Administration, St. Helens Station.

2.2.2 Geology

Regionally, the Site lies in the Puget-Willamette Lowland, a broad structural and topographic depression located between the Coast Range to the west and the Cascade Range to the east. St. Helens lies in the northern portion of the Portland Basin, one of several complexly faulted basins comprising the Willamette Valley segment of the Puget-Willamette Lowland. The Portland Basin is interpreted to be a pull-apart basin between two *en echelon* right-lateral strike-slip faults (Evarts, 2004) caused by regional compression and shear associated with oblique subduction of the Juan de Fuca Plate beneath the North American Plate off the Oregon coast. In the deepest part of the Portland Basin, up to 1,800 feet of sediment has accumulated since the late Miocene.

The Site is on the west side of Multnomah Channel. The topographic highlands to the north expose Miocene flood basalts of the Grand Ronde Formation, deeply weathered in places. Chemistry and magnetostratigraphy data indicate the basalt is equivalent to the Member of Sentinel Bluffs, which is dated to approximately 15.6 million years. In the area of the Site, the Multnomah Channel is separated from the Columbia River by a peninsula marking the northernmost extent of Sauvie Island. The islands, floodplains and point bars of Multnomah Channel and the Columbia River east of the Site are comprised of Quaternary alluvium (i.e., primarily fine sand and silt).

Soil encountered at the Site consists of fill underlain by native soil, which in turn is underlain by basalt bedrock. The fill ranges from 4 to 12 feet below ground surface (bgs) and consists predominantly of dredge sand with localized areas of silt, coarse gravel, cobbles, wood chips, and metal and brick debris. The native soil, beneath the fill, ranges from 4 to 24 feet bgs and consists primarily of silt with varying amounts of clay and organic matter, and minor amounts of sand and gravel. Basalt bedrock was generally encountered at depths between 6 and 35 feet bgs in the upland area and at depths of 2 to 26 feet beneath river sediment in the offshore area. Bedrock beneath the Site is part of the Grande Ronde flows of the Columbia River Basalt. Detailed cross-sections were prepared for transects through the upland and in-water areas shown on Figure 19. The crosssections are presented as Figures 20 through 27. The top of basalt surface elevation contours for the upland and in-water area are shown on Figure 28.

2.2.3 Hydrogeology

Regionally, shallow groundwater occurs in the surface fill materials, alluvium, and where the basalt bedrock is near the surface, in the weathered upper flow surfaces. Deeper groundwater zones are located in the more permeable interflow zones between the unweathered basalt flows and are not hydraulically connected to shallow groundwater.

In 1996, 17 monitoring wells, 13 located in Area 1, were installed as part of the risk investigation (Figure 29). Five monitoring wells were installed as pairs (depicted by A/B) with one deeper, though located in same water bearing zone, than the other for a total of 22 monitoring wells. One additional monitoring well was installed in 1998. The monitoring wells were constructed to depths ranging from 9 to 20 feet bgs depending on when bedrock was encountered. Well screens within the 2-inch diameter polyvinyl chloride (PVC) well casings ranged from 5 feet in length to 10 feet.

Following installation and development, depth to water measurements were regularly collected from the monitoring wells from August 1996 to March 1999. Groundwater was measured at depths

ranging from 2 to 8.5 feet bgs and fluctuates seasonally. Groundwater generally flows towards the nearest water body. Groundwater discharge to surface water is strongly influenced by seasonal rainfall and fluctuations in river stage. Groundwater contours for high and low water conditions are shown on Figure 29.

2.2.4 Surface Water and Stormwater

The Site lies on the northwestern bank of Scappoose Bay near the confluence of Scappoose Bay and the Multnomah Channel of the Willamette River. The confluence of the Multnomah Channel and the Columbia River is approximately 1.3 miles downriver from the Site. Milton Creek is a perennial stream located on the Site's western boundary and discharges to Scappoose Bay at the southwest corner of the Site. The Upper Milton Creek PAA is located in and adjacent to Milton Creek. The main channel of Scappoose Bay is offshore of Area 1 Upland. The Cove Area and Area 1 Dock PAAs are located in Scappoose Bay. Area 2 Dock PAA is located in the Multnomah Channel, downstream of Scappoose Bay. The locations of the PAAs and surface water bodies are shown on Figure 2. Bathymetry contours from 2010 and 2017 surveys are shown on Figures 30 and 31.

Based on a review of bathymetric maps generated in 2010 and 2017, the main channel of Scappoose Bay is a stable feature in the vicinity of Area 1. Tidal fluctuations of water levels occur daily in Scappoose Bay and Multnomah Channel, in addition to seasonal fluctuations in water levels as a result of runoff events in the watershed. The Site's shoreline is routinely inundated above the OHWM (14 feet NAVD88) during high tides (typically between April and July). Mudflats are commonly visible along the Site's shoreline between approximately 5 to 10 feet NAVD88. Depending on shoreline conditions, basalt topography, and channel dimensions, portions of the Site's shoreline experience sediment deposition (e.g., flat, soft bottom cove areas), while riverbank areas closer to the main channel appear to exhibit localized erosion of sediment from large storm events and/or tidal cycles. Specific conditions of each in-water PAA are described below.

2.2.4.1 Area 1 Dock

The 700-foot-long Area 1 Dock shoreline and in-water area is closer to the main channel of Scappoose Bay and is subject to stronger river current. The main channel of Scappoose Bay widens and deepens adjacent to and within the Area 1 Dock. Specifically, the bottom of the main channel ranges between 200 and 250 feet NAVD88 wide and between -10 and -5 feet NAVD88 deep. The river bottom slopes steeply downward within 50 to 100 feet of the shoreline. The Area 1 Dock shoreline is relatively stable to eroding, which generally prevents the deposition of newer clean sediments and the burial of historical contamination.

Deteriorating creosote-treated timber pilings, remnants of dock structures, are present throughout this area. Finer textured wood debris, described as "pulverized wood" in sampling logs and is interpreted to be hog fuel (e.g., wood shavings/chips) likely spilled from the hog fuel conveyor, is also present in sediment. The distribution and thickness of the woody debris is highly variable in this area, ranging from completely absent to more than 11 feet thick.

Contaminated sediments in the Area 1 Dock have been characterized to an approximate depth of -25 NAVD88 or about 25 feet bml. Sediments in the western portion of the area, in the vicinity of the hog fuel conveyor and loading hopper, were characterized as sand to sandy silt. Sediment in the eastern portion of the area, in the vicinity of the transfer table and dock, were characterized as silt. Basalt bedrock was encountered at various depths throughout the Area 1 Dock. Crosssections along transects B-B', G-G', and H-H' (see Figure 19 for cross-section locations) are included as Figures 21, 26, and 27.

2.2.4.2 Area 2 Dock

The 600-foot long and 150-foot-wide Area 2 Dock shoreline and in-water area is in the Multnomah Channel. Bathymetry data is not available for this area. Only a limited bml investigation of the Area 2 Dock has been performed. Surface sediments generally consist of silty sand with varying amounts of creosote wood debris.

Based on the presence of contamination in surface sediment and the hard compact nature of the mudflats with relatively steep slopes towards the channel, the Area 2 Dock shoreline appears to be relatively stable to eroding. Creosote-treated timber pilings, remnants of dock structures, define this area. The main stem of the Multnomah Channel is 100 to 150 feet offshore with a bottom elevation of approximately -15 feet NAVD88 immediately downstream.

2.2.4.3 Cove Area

Prior to development, the Cove Area consisted of a partially vegetated slough. The Cove Area was created between 1948 and 1953 through extensive dredging and filling of historical lowlands (see Figures 7, 8, and 9). The Cove Area is a depositional area, protected from erosional effects of large storm and tidal surges.

The elevation of the river bottom within the cove and shallow inlet immediately downriver of the peninsula ranges between approximately 0 and 10 feet NAVD88 within approximately 100 to 300 feet of the shoreline before sloping steeply towards the main channel of Scappoose Bay. The main channel of Scappoose Bay adjacent to the Cove Area widens to approximately 150 feet with bottom elevations between -7 and -5 feet NAVD88.

After operations ceased in 1960, deposition and burial of the old operational sediment surface occurred throughout the Cove Area. A sediment deposition rate of 0.2 to 1.4 in/yr is implied based on the presence of 1 to 7 feet of fine textured sediment above the old contaminated operational sediment surface characterized by creosote wood debris.

The buried layer of creosote wood debris is interpreted to be the former in-water surface during plant operation. The creosote wood debris ranges from 1 to 7 feet bml beneath much of the Cove Area. The thickness of the creosote wood debris progressively thins and diminishes in magnitude further from the shoreline.

Contaminated sediments in the Cove Area have been characterized to an approximate depth of -25 NAVD88 or about 25 feet bml. Sediments were characterized as silt. Basalt bedrock was encountered at various depths through the Cove Area. Cross-sections along transects A-A', D-D', F-F', and G-G' (see Figure 19 for cross-section locations) are included as Figures 20, 23, 25, and 26.

2.2.4.4 Upper Milton Creek

Upper Milton Creek's sediment profile consists of up to 1 foot of silty organic-rich material overlying a 1- to 6-foot layer of gray, stiff, clayey-sandy silt with gravel and cobbles. Scouring of the creek channel by seasonal runoff and tidal cycles limits accumulation of sediment in this portion of the creek. The shoreline is generally a relatively steep embankment covered with vegetation.

2.2.4.5 Stormwater

No stormwater management features were identified during a review of historical facility maps.

2.3 PLANT OPERATIONS

The Site was first developed for industrial purposes in 1912 when St. Helens Creosoting Company established a wood treating facility. A sawmill operated in the northeast portion of the Site from 1915 to the mid-1930s (see Figure 3B). Companies acquired by Pope and Talbot, Inc. purchased the property in 1938 and continued to operate the wood treating facility until 1960. The plant was dismantled in the fall of 1960. The Port of St. Helens (now the Port of Columbia County) purchased the property in 1963.

The Site was vacant between 1960 and 1974. Dredge material generated from deepening of the Columbia River was placed on the property in the early 1970s in an effort to level and raise the property's surface grade by several feet (see Figure 12). Since 1974, the Site has either been vacant or leased for the following industrial uses:

- Pole peeling and pole storage facility (1974 to 1991).
- Storage yard for a marine construction/dredging company (1993 to 1998).
- Small private wood-working business (2000 to 2005).
- Marine log salvage and sawmill (2006 to 2008).
- Log storage, pole peeling, and sawmill (2009 to 2012).
- Miscellaneous small-scale private businesses (2014 to present).

Figures showing facility features during different periods are included as Figures 4 through 16.

2.3.1 Physical Plant

While multiple companies have operated at the Site, there have been two primary physical plant layouts.

2.3.1.1 Wood-Treating Facility

The original wood-treating facility constructed in 1912 consisted of a creosote AST, pipeline, a retort/pump house, creosote retorts, transfer table, tram tracks, and dock. The mill consisted of a mill building, power house, and vent burner. By 1929 the wood-treating facility had expanded to include three large volume creosote ASTs, with removal of the original creosote AST, two fuel oil ASTs, a crane fuel AST, blacksmith/machine shop, boiler house, experimental kilns, hog fuel bin, conveyor and hopper, a barge house, and additional docks. An above and below ground pipeline conveyed creosote from the largest creosote AST to the eastern dock (Area 2 Dock). This facility configuration essentially remained until the plant was dismantled in 1960. Figures 3A and 3B shows historical features.

Creosote was delivered to the Site via ships which transferred their contents at the Area 2 dock. An aboveground pipeline pumped the creosote to a large AST from which creosote was pumped to several smaller ASTs. Open "process recovery tanks" collected condensed water from the Boulton treatment process. This process consists of enclosing wood in a treatment cylinder, introducing hot creosote or pentachlorophenol (PCP) solution and establishing a vacuum at 220°F in the cylinder. This process removes water from the wood as water vapor that is condensed outside the cylinder. The collected water then passed through a series of chambers with skimmers to recover any treatment solution with filtered process water discharging to Scappoose Bay (See Figure 3A). All other waste materials were recycled.

2.3.1.2 Post-Peeling Facility

An office and maintenance shop were built in 1974 or 1975 for the pole-peeling operation. Two underground storage tanks (USTs) and two waste oil ASTs were present. The 10,000-gallon diesel fuel UST and the 1,000-gallon gasoline UST were installed in 1975 adjacent to the maintenance shop building. Both USTs were decommissioned in-place in 1990. DEQ assigned UST Facility No. 5196 to the USTs. Figures 3A and 3B shows historical features.

2.3.2 Chemical Use and Waste Generation and Management

The follow section discusses the available information for the two primary operations that were present on the Site.

2.3.2.1 Wood-Treating Facility

Products that were treated included pilings, poles, cross arms, paving blocks, pipe and tank staves, structural timbers, and railroad cross ties. The primary wood-treating formula used was creosote. PCP was reportedly used in 1953 and 1954. Chromated copper arsenate, a wolmanizing salt, was reportedly used in 1953. No other information on chemical use and waste generation and management from the wood-treating facility has been located.

2.4 REGULATORY HISTORY

Investigations have been performed at the Site and adjacent in-water areas since 1988. The investigations are documented in the administrative file (ECSI No. 959). Information is also available online through DEQ's ECSI database.

2.4.1 Preliminary Assessment (1988) and Site Inspection (1990)

DEQ performed a Preliminary Assessment (PA) of the Site in 1988. The PA identified unanswered questions related to past site uses and disposal practices that may have impacts on sensitive environments and requested the U.S. Environmental Protection Agency (EPA) perform a Site Inspection.

E&E, on behalf of the EPA, performed a Site Inspection from July 1989 to May 1990. The Site Inspection was performed to further assess potential source areas, evaluate contaminant migration pathways and identify areas warranting additional investigation. E&E collected eight subsurface soil samples, one surface soil sample, three seep solid samples, four surface water samples, and five sediment samples. Additional samples were collected off-Site for background purposes. Sample locations are shown on Figure 32. Samples were analyzed for one or more of the following: volatile organic compounds (VOCs), semi-volatile organic compounds (SVOCs), pesticides, polychlorinated biphenyls (PCBs), phenols, and metals.

PAHs associated with creosote were detected in subsurface soil, seep solid, sediment, and surface water samples collected at or immediately downgradient of the former wood treating facilities. PCP was detected in one subsurface soil sample collected near the former PCP and wolmanizing salts retorts.

2.4.2 Site Investigation (1993)

Harding Lawson Associates (HLA) performed a site investigation above the groundwater table from March to April 1993. The site investigation was performed to further assess potential source areas. HLA advanced 32 soil borings, the majority located in Area 1. A total of 51 subsurface soil samples, five surface soil sample, and one seep solid samples were analyzed. Sample locations are shown on Figure 32.

Samples were analyzed for contaminants including gasoline, diesel, heavy oil, VOCs, PAHs, and metals. Gasoline, diesel, heavy oil, VOCs, PAHs (mainly PAHs and PCP), and/or metals were detected in subsurface soil samples. The site investigation concluded the majority of contaminated soil is present in the vicinity of the former wood treating facilities and in the native soil located below the dredge fill material. Concentrations of arsenic were determined to be typical of background with the possible exception of arsenic detected in immediate vicinity of the former wood treatment building.

2.4.3 1995 Consent Order & 2008 Orphan Declaration

DEQ issued an Order on Consent (Consent Order) to Pope and Talbot, Inc. and the Port of St. Helens (now Port of Columbia County). The Consent Order required the parties to conduct a Remedial Investigation and prepare a Feasibility Study.

Pope and Talbot, Inc. declared bankruptcy in November 2007 and ceased performing work under the Consent Order. The Port of St. Helens ceased work shortly after due to financial constraints. DEQ declared the Site an Orphan on April 28, 2008. The Port of St. Helens later identified insurance policies and tendered claims to the insurance companies. The Port of St. Helens then continued the work required under the Consent Order. The insurance policies are expected to cover a portion of the selected remedial action but not the entire cost.

2.4.4 2000 Remedial Investigation

Investigations were performed at the Site in two phases between 1996 and 1998. This investigation included both Area 1 and Area 2. The results of the investigations are included in the April 7, 2000 Remedial Investigation report prepared by GeoEngineers, Inc. The investigations were performed to evaluate the nature, extent, and risk associated with historic releases of hazardous substances at the Site. A summary of samples collected, analysis performed, and matrices sampled throughout the remedial investigation period (1996 through 2017) is provided on Table 1.

2.4.4.1 Phase 1 Investigation (1996 to 1997)

The initial phase of investigation was performed between July 1996 and November 1997. The investigation consisted of upland soil and groundwater and in-water sediment and surface water immediately adjacent to the Site's shorelines. A total of 31 soil borings were advanced, generally until basalt bedrock was encountered, at the Site in July 1996. Monitoring wells were constructed in 22 of the soil borings. Boring and well locations are shown on Figure 32.

Groundwater elevation measurements were collected on a monthly basis with groundwater samples collected quarterly. Six groundwater monitoring events were performed during Phase 1.

Surface water (SW1 through SW4) and sediment samples (SHI01S, SHI01B through SHI12B) were collected adjacent to the Site in Milton Creek, Scappoose Bay, and Multnomah Channel in October and November 1996. Surface water and sediment sample locations are shown on Figure 32.

2.4.4.2 Phase 2 Investigation (1998)

Additional investigation of surface soil, subsurface soil, groundwater, riverbank seeps, surface water, and bedrock was performed in 1998. A total of 26 soil borings (GP1 through GP26) were advanced at the Site in September 1998. One additional monitoring well (MW-18) was constructed. Three borings (RCB-1, RCB-2, RCB-3) were advanced into the basalt to evaluate hydraulic characteristics of the basalt formation. Boring and well locations are shown on Figure 33.

Groundwater from monitoring wells, surface water (SW-MC1, SW-MC2, SW-SB1 through SW-SB7), and a surface soil sample (Bank1) near a seep were collected in September 1998. Seep water samples (Seep1, Seep2, Seep3) were collected in October 1998. Surface samples (SS1 through SS24) were also collected to evaluate the gravel roadways in October 1998. Soil sample and seep locations are shown on Figure 33.

2.4.4.3 Phase 1 & 2 Investigation Results

A visible creosote-petroleum product (NAPL) was observed in subsurface soil samples and in groundwater monitoring wells in Area 1 Upland. In borings completed within close proximity of the former wood-treating operations, creosote impacts were observed throughout the entire thickness of the native soil unit (depths ranging between about 5 to 18 feet bgs). The vertical thickness of creosote-impacted soil appears to decrease with distance from the wood treatment plant and ASTs.

All media samples were submitted for chemical analysis for one or more of the following contaminants of potential concern (COPC): total petroleum hydrocarbons (TPH); VOCs; SVOCs, chlorinated phenols; dioxins and furans; and select metals (arsenic, cadmium, chromium, copper, lead, and zinc). Sediment samples were also analyzed for grain size, sulfides, and total organic carbon (TOC). Bioassay tests were performed on two sediment samples.

In general, TPH, VOCs, and SVOCs associated with creosote were detected in groundwater samples obtained from monitoring wells and soil borings completed in the vicinity of the former wood treating operations. Benzene, toluene, ethylbenzene, and xylenes (BTEX) represents the majority of VOCs detected in groundwater samples. PAHs represent the majority of the SVOCs detected in groundwater samples. Relatively low levels of metals and chlorinated phenols were detected in Area 1 Upland groundwater samples.

During periods of low surface water, small, localized groundwater seeps have been observed along the steeper sections of the Scappoose Bay and Milton Creek shoreline adjacent to Area 1 Upland. Many of these seeps appear to daylight near the fill and native soil interface. Creosote-impacted seeps occasionally are accompanied by a sheen or visible NAPL, although at times the sheen has been a natural (organic) sheen not related to creosote. BTEX and PAHs were detected in October 1998 seep water samples obtained from the Scappoose Bay and Milton Creek shorelines.

Shallow sediment conditions (upper 10 centimeters) along the Area 1 Upland shoreline at the Site were explored by obtaining a series of discrete and composite sediment samples along eight transects and background locations. PAHs were detected in the sediment samples.

Subsequent investigations were conducted to further evaluate the nature and extent of contaminants in sediments in preparation for completion of human health and ecological risk assessments.

2.4.5 Interim Remedial Investigations (1999-2006)

This section discusses actions conducted after the data set collected for the 2000 Remedial Investigation report and before the start of the supplementation remedial investigation data collection.

Annual groundwater monitoring was conducted in 1999, 2000, 2001, 2003, 2004, 2005, and 2006. Groundwater samples were analyzed for TPH, BTEX, PAHs, and dissolved metals (arsenic, cadmium, chromium, copper, lead, and zinc).

Sediment sampling was completed in 2003 (SD-100 through SD-120 and SD-BG-01 through SD-BG05), 2004 (SD-103, SD-104, SD-105, SD-112, SD-117, SD-119, SD-121 to SD-124, SD-MC-A through SD-MC-E), and 2005 (SD-125 through SD-141) as part of an ecological risk assessment (ERA). Offshore surface sediment samples obtained adjacent to the Site and from background locations were analyzed for one or more of the following constituents: PAHs, arsenic, dioxin and furans, TOC, total sulfides, ammonia, grain size distribution, and bioassay testing. Sample locations are shown on Figure 32.

A baseline human health risk assessment (HHRA) and ERA was completed in 2006. The results of the HHRA and ERA are discussed in Sections 3.2.3 and 3.2.4, respectively.

2.4.6 Supplemental Remedial Investigation (2010 – 2018)

The 2000 Remedial Investigation Report presented data collected at the Site through 1998. The initial phases of the remedial investigation were focused on the upland portions of the Site, with only limited characterization of surface water and surface sediment along its 4,500 linear feet of shoreline of Milton Creek, Scappoose Bay, and the Multnomah Channel. Unlike the site-wide assessment described in the 2000 Remedial Investigation Report, the supplemental remedial investigation findings are focused on the Area 1 Upland, offshore areas adjacent to Area 1 within Milton Creek and Scappoose Bay, and the offshore area adjacent to Area 2 within Multnomah Channel near the terminus of the former creosote pipeline and residual pilings associated with a historical dock. Multiple investigations were performed between 2010 through 2018 as part of the supplemental remedial investigation. The actions performed are summarized below. Sample locations are shown on Figure 33.

2.4.6.1 Investigations Performed

Quarterly groundwater monitoring of Area 1 wells was performed between September 2010 and August 2012. Groundwater samples were analyzed for TPH, BTEX, PAHs, and dissolved metals.

Multiple phases of shoreline inspections were conducted during low water conditions (generally in late September and early October) in 2010, 2011, 2012, 2017, and 2018 to designate specific areas where NAPL (i.e., creosote sheen, product, and/or tar body) was observed in near-surface sediment, and to refine the locations and boundaries of visible creosote-impacted groundwater seeps in the bank. Surface sediment samples were collected at areas of obvious or suspected creosote impact and were analyzed for TOC, TPH, PAHs, VOCs, extractable petroleum hydrocarbon (EPH) and volatile petroleum hydrocarbon (VPH) constituents, and total metals (arsenic, cadmium, chromium, copper, lead, and zinc).

Two phases of upland and offshore NAPL delineation were conducted in 2010 and 2011. The initial delineation in 2010, generally consisted of soil boring equipment equipped with TarGOST® laser-induced fluorescence technology. The second delineation effort consisted of direct observations of sampling cores and sediment, soil, and basalt sample collection for analysis from borings. Soil samples were submitted for TPH and PAHs analysis. Sediment samples were analyzed for TPH, PAHs, TOC, and grain size.

Additional monitoring wells (MW-2A, MW-19, MW-20, MW-21, MW-22, MW-23, MW-24, and MW-25) were installed in 2011 within Area 1 Upland to better delineate the upland extent of creosote NAPL. Groundwater monitoring was performed in July 2011.

In 2012, sediment and porewater sampling was conducted adjacent to Area 1 in Scappoose Bay and Milton Creek to assess the bioavailable or bioaccessible concentrations of PAHs associated with creosote within and near the base of the benthic (biologically active) sediment environment.

To evaluate potential temporal variability resulting from changes in river stage and groundwater elevation, porewater sampling was conducted in July 2012 (high river stage and high groundwater elevation) and October 2012 (low river stage and low groundwater elevation). Whole or bulk sediment samples were analyzed for 18 parent PAH compounds and 16 groups of alkylated PAH homologues, TOC, and black carbon. Sediment porewater samples were analyzed for 34 PAHs (parent and alkylated PAHs)

In August 2013, the dilapidated wooden deck of the former transfer table dock was removed to facilitate the October 2013 delineation of creosote-related contamination within the underlying nearshore soil and sediment. Soil and sediment samples from varying depths (dependent upon field observations) were visually screened for creosote NAPL and submitted for TPH and PAHs analysis.

A phased offshore data gap investigation consisting of a high-accuracy bathymetric survey, offshore groundwater discharge survey, sheen mapping in surface sediment, and in situ porewater and surface water passive sampling was completed between April 2017 and November 2018, to further assess the offshore extent of creosote NAPL and the risks that bioavailable fractions of PAHs and TPH pose to aquatic invertebrate receptors outside the inferred NAPL areas. Surface sediment samples were analyzed for 34 PAHs, VPH, EPH, VOCs, TOC, and black carbon. Time integrated (20 days) surface water (1 foot above mudline) and sediment porewater (two depth intervals of 3 to 8 inches and 24 to 29 inches bml) were sampled using passive sampling devices and analyzed for freely dissolved PAHs (34 and 63 compounds), VPH, oxygenated PAHs (OPAHs), and hopane and cholestane compounds.

2.4.6.2 Investigation Results

The conceptual site model (CSM) was refined to better account for offshore sediment contamination (particularly NAPL along the Site's shorelines), as well as upland NAPL migration and dissolved-phase contaminant transport. A schematic CSM is included as Figure 34.

The lateral and vertical extent of NAPL in offshore sediment adjacent to the Site was defined. Wood debris in surface sediment along 700 feet of the Area 1 shoreline is a source of NAPL beneath the former hog fuel dock and transfer table dock structures and extends approximately 75 feet offshore. The creosote contaminated wood debris beneath the former transfer table dock extends approximately 200 feet offshore.

Although upland groundwater discharges to sediment porewater and surface water within Milton Creek and Scappoose Bay, the risk to human health and ecological receptors from dissolved-phase constituents transported via groundwater appears to be low outside the observed areas of direct NAPL seepage.

The updated supplemental HHRA indicated low potential risk to human receptors other than direct contact with NAPL and Area 1 groundwater by excavation workers and direct contact with NAPL by shoreline anglers and/or transients.

The updated supplemental ERA indicates that the presence of creosote NAPL (inclusive of moderate to heavy petroleum sheen) presents unacceptable ecological risk to benthic invertebrates and fish where it is observed in surface sediment. In general, the 2017 surface sediment porewater sampling indicated dissolved-phase constituents in deeper sediment and/or upland groundwater do not adversely impact benthic invertebrates outside the prior remedial action areas.

The hot spot evaluation indicated that groundwater and sediment within the delineated extent of NAPL and NAPL-affected groundwater seeps are hot spots in accordance with DEQ guidance. Dissolved-phase creosote constituents in porewater and surface water that cause unacceptable risk to benthos and fish are also potential hot spots.

3.1 NATURE AND EXTENT OF CONTAMINATION

As summarized previously, investigations of the Site commenced in 1988 and continued through 2018. While the early investigations (e.g., 1988 PA, 1990 Site Inspection, 1993 Site Investigation, 2000 Remedial Investigation) included sampling in the entire Upland Area, the Supplemental Remedial Investigation focused primarily on Area 1 Upland, offshore areas adjacent to Area 1 within Milton Creek and Scappoose Bay, and the offshore area adjacent to Area 2 within Multnomah Channel near the terminus of the former creosote pipeline and residual pilings associated with a historical dock. The results of the Supplemental Remedial Investigation included the identification of five PAAs by contaminants including creosote (TPH) and PAHs. The nature and extent of contaminants are summarized below.

The Locality of Facility (LOF) is defined as any point where a human or ecologic receptor contacts or is reasonably likely to come into contact with contaminants from the Site. The LOF takes into account the likelihood of the contaminants migrating over time. For the purpose of this ROD, the LOF is the soil and shallow water-bearing zone above Basalt Bedrock within Area 1 Upland and portions of sediment and surface water in Milton Creek and Scappoose Bay. The extent of the LOF is shown on Figure 2.

3.1.1 Non-Aqueous Phase Liquid

Creosote NAPL has been encountered in soil, groundwater, groundwater seeps, and sediment at the Site. The NAPL composition is dominated by diesel- and oil-range hydrocarbons (although chromatograms do not match a diesel or oil fingerprint) and PAHs (naphthalene, phenanthrene, 2 methylnaphthalene, and others). The majority of creosote NAPL appears to have been trapped by capillary forces within the native soil before reaching the basalt bedrock. The total thickness of creosote impact varies from 1 to 8 feet, and the impacted intervals typically occur at depths greater than 10 feet bgs (below the Fill Zone). The estimated lateral extent of NAPL and creosote sheens are shown on Figures 35, 36, and 37.

3.1.1.1 Area 1 Upland Non-Aqueous Phase Liquid

The Area 1 Upland extent of NAPL, although not continuous throughout, has been generally defined within an approximately 4-acre area. The distribution of NAPL beneath the former operations area appears to be a function of subsurface migration from the former creosote storage tanks, plant, and retorts as well as direct releases from other ancillary operations (e.g., the transfer table).

Potentially mobile creosote is present in the native soil at various depths below the fill and above the basalt bedrock and within distinct stratigraphic layers that are less than 18 inches thick. The total thickness of creosote varies from 1 to 8 feet. In general, this vertical layering or profile of NAPL becomes thinner and deeper within the native soil farther from the former operations area at depths greater than 10 feet bgs.

The inferred extent of NAPL indicates that topographical highs on the basalt surface, such as the ridge to the east of the former operations area along the Scappoose Bay shoreline, appear to influence the subsurface distribution of NAPL. The bedrock appears to be a barrier to downward vertical migration of the NAPL.

3.1.1.2 Area 1 Dock Non-Aqueous Phase Liquid

Between the former hog fuel dock and northeastern (downriver) terminus of the former transfer table dock (Area 1 Dock), discontinuous, isolated pockets of creosote NAPL have been observed in many locations at or near the sediment surface.

The source of sediment contamination is creosote contaminated and/or treated wood debris. This debris differs in character from debris found in the Cove Area. Much of the wood debris is finer textured and is described as "pulverized wood" in sediment sampling logs. The pulverized wood is interpreted to be hog fuel (i.e., wood shavings/chips) likely spilled from the hog fuel conveyor and during over-water transfers from ships/barges to the hog fuel hopper and conveyor system. Some of this material appears to have become saturated with creosote and accumulated along the shoreline and in sediment surrounding, and downstream of, the former hog fuel dock.

Additionally, some of the contaminated wood debris appears to be related to deterioration of the creosote-treated dock structures (i.e., timber piles, lateral bracing, decking) that existed along this portion of the Area 1 shoreline. The distribution and thickness of the woody debris is highly variable in this portion of the Site, ranging from completely absent to more than 11 feet thick (sediment boring SB-16). The lateral extent of surface sediment and/or wood debris capable of producing a creosote-related sheen on the overlying water surface when disturbed extends approximately 75 feet offshore between the former hog fuel dock and transfer table dock structure.

Within the western half of the primary wood dock structure of the former transfer table, thick sequences of contaminated wood debris are intermingled with dimensional lumber, logs, metal ties, and other debris. Heavy sheens are easily generated by disturbing the surface sediment around many of the remaining log pilings.

Creosote NAPL associated with wood debris was primarily observed in the upper 24 inches of sediment within the central and western portions of the former transfer table dock area. Deposited wood and creosote-related contamination were observed from the surface to approximately 9 feet bml in the far western portion of the former transfer table dock. In general, the thickness of the wood mass decreased to the northeast away from the southwest area of the former transfer table dock and is not present in significant thickness to the north of boring DAI-HA-02 or to the east of location DAI-PH-11. The lateral extent of surface sediment creosote contaminated wood debris extends approximately 200 feet offshore within the footprint of the former transfer table dock.

An upland and slightly in-water area of subsurface creosote staining and isolated blebs of NAPL were encountered in the native soil beneath the northwest corner of the former transfer table dock. Creosote impacted soil with residual NAPL decreases in thickness and is encountered at deeper intervals within the native soil as distance increases in all directions from the northwest corner of the former transfer table dock. This corner of the former transfer table dock is covered by recently deposited sand (natural deposition) and there is no visible surface evidence of impacted groundwater seeps from this upland area of NAPL.

3.1.1.3 Area 2 Dock NAPL

In October 2017, DEQ staff inspected the Area 2 shoreline and observed creosote-impacted surface sediment near the terminus of the former creosote pipeline and residual piling associated with a historical dock (Area 2 Dock). Creosote sheen occurrence was mapped in surface sediment along the approximately 900-foot-long portion of the Scappoose Bay shoreline. Moderate to heavy petroleum sheens were noted only in borings amongst the pilings near the shoreline that encountered creosote-contaminated wood debris within the upper 12 inches of sediment. Sediment poling beyond 50 feet of the shoreline (even amongst the pilings), did not produce a visible sheen. The lateral extent of creosote contaminated wood debris covers an approximately 200-foot-long by 50-foot-wide section (0.2 acre) of the Area 2 shoreline beneath and immediately downriver of the former dock and creosote AST.

3.1.1.4 Cove Area NAPL

Creosote and fuel oil ASTs were immediately upland of the Cove Area and a hog fuel boiler and pole peeling station operated along the shoreline. Based on historical photographs and information on plant operations and its subsequent demolition, significant amounts of creosote contaminated hog fuel, tree bark, and dimensional lumber appear to have accumulated in the Cove Area. NAPL has migrated from these upland source areas into approximately 0.8 acre of the Cove Area.

Heavy petroleum sheens are visible in seasonal groundwater seeps along a 300-foot section of bank seepage west of the peninsula. Near-shore Cove Area borings typically encountered 6 to 8 feet of NAPL impacted sediment, with a maximum thickness of 12 feet observed in SB-11. The near-shore presence of moderate to heavy petroleum sheens in surface sediment appears to be related to the NAPL/groundwater seeps, the presence of creosote-treated wood debris, and gas ebullition resulting from biodegradation of organic matter in deeper sediments.

Moderate to heavy petroleum sheens have been detected in subsurface sediment up to 220 feet offshore beneath the Cove Area (borings SB-04 and SB-09). In general, creosote-impacted sediment decreases in thickness and is encountered at deeper intervals within the Cove Area as distance increases from the shoreline. Outside the inferred extent of upland NAPL, the vertical mobilization of sheen via ebullition has not been observed in the Cove Area.

Riverward of the solid red NAPL boundary depicted on Figure 37, the source of subsurface sediment contamination transitions to creosote contaminated wood debris. This buried layer of creosote wood debris is interpreted to be the old in-water operational surface while the plant was active and during its demolition. Creosote-impacted wood debris ranges from 1 to 7 feet bml beneath much of the Cove Area, and progressively thins and diminishes in magnitude outward from the shoreline. Vertically, contaminant levels attenuate rapidly within the sediment stratigraphic layers. Sediment within the western portion of the Cove Area is generally unimpacted by upland NAPL migration and/or creosote contaminated wood debris.

Light or slight petroleum sheens have been intermittently observed in the relatively clean, more recently deposited surface sediment beyond the solid NAPL boundary depicted on Figure 37. Based on the results of bulk sediment and depth-discrete porewater sampling, these slight surficial sheens appear to be attributed to localized particles of creosote-treated wood debris in the surface sediment and do not appear to be attributed to NAPL migration from the underlying sediment.

3.1.1.5 Upper Milton Creek Area NAPL

Groundwater seepage with a creosote-related sheen has been observed along a 200-foot section of the east bank of the Upper Milton Creek and analytical results of bank samples indicate contact with NAPL. The Upper Milton Creek in-water area is located over 350 feet west of any known historical wood-treating features and/or operations. Field evidence of NAPL was generally not observed in borings.

The primary sources of contamination to the Upper Milton Creek in-water area are heavy petroleum sheens seeping from a relatively thin (1-foot-thick or less) sand layer situated at an approximate depth of 19 to 20 feet below the top of the bank. Biogenic sheens also seep from the eastern bank of Milton Creek and are intermingled with petroleum sheens. The platy relatively odorless biogenic sheens, which are easily distinguished from the fluid, odoriferous creosote related petroleum sheens likely emanate from areas where upland creosote contamination and wood debris is more completely degraded.

3.1.2 Groundwater

Groundwater monitoring has been performed frequently enough to provide understanding of groundwater conditions. Residual NAPL continues to be a source for dissolved-phase creosoterelated constituents in shallow groundwater beneath the Site. The extent of PAH and TPH groundwater contamination is shown on Figures 38 and 39, respectively.

The highest concentrations of PAHs and TPH in groundwater were generally detected in monitoring wells MW 22, MW-2A, and MW-3A, within the inferred NAPL area, followed by detections in wells at the southeastern and northeastern edges (MW-23, MW-24, and MW-25) of the former wood treating operations area. Detections of PAHs were generally lower in wells to the west of and outside of the inferred upland NAPL area and in wells at the far eastern edge of Area 1. Analytical results from groundwater samples collected during the supplemental remedial investigation are representative of current Site conditions and are summarized below.

Total PAHs were detected in groundwater samples collected within or near the former operations area, with a similar distribution of relative concentrations as TPH. Total PAHs detected in groundwater ranged in concentration from 0.35 micrograms per liter $(\mu g/L)$ at well MW-18 to a maximum of approximately 26,800 μg/L at well MW-3A in 2010; NAPL was also detected at MW-3A during the 2010 monitoring event. The maximum total PAHs in groundwater in 2011/2012 was approximately 22,500 μg/L in well MW-22.

Petroleum hydrocarbons indicative of creosote were detected in groundwater samples collected within or near the former operations area. TPH detections were primarily in the diesel range. The maximum diesel-range hydrocarbon concentration was detected in well MW-3A at 374 milligrams per liter (mg/L) in 2010. NAPL was also detected at MW-3A and sampled during the 2010 monitoring event. The maximum diesel-range hydrocarbon concentration in 2011/2012 was 116 mg/L at MW-3A. The maximum concentration of TPH in the oil range was 103 mg/L at MW-3A in 2010; the maximum oil-range hydrocarbon concentration in 2011/2012 was 3.81 mg/L at MW-2A.

Metals detected in groundwater were arsenic, chromium, copper, lead, and zinc. The distribution of metal detections does not generally correlate with NAPL or the former wood treating operations area. Arsenic was detected in every well sampled at concentrations up to 17.5 μg/L at MW-23. Chromium was detected infrequently just above reporting limits in MW-3B and MW-5, and at slightly higher concentrations (maximum of 5.83 μg/L) at MW-2A. Copper was detected at MW-5 at a maximum concentration of 18.5 μg/L; the only other detection of copper was just above the reporting limit at MW-4B. Lead was detected infrequently in wells MW-3B, MW-5, and MW-9, at levels up to 3.48 μg/L. The maximum zinc detection, 683 μg/L, was at MW-20 in 2011; other detections at MW-20 were below 10 μg/L. Well MW-5, located near the northern property boundary, has had consistent detections above 200 μg/L. All other zinc concentrations were below 200 μg/L, with most below 10 μg/L.

3.1.3 Soil

The fill and native soil contact generally represents the original ground surface at the time of woodtreating facility operation. This results in the creosote-impacted soil in the vicinity of the former wood-treating facilities being generally limited to the native soil unit, located beneath 5 to 10 feet of dredge sand fill. In borings completed in close proximity to the former wood-treating plant and former creosote ASTs, varying degrees of creosote impact have been observed throughout the entire thickness of the native soil unit with depths ranging between about 5 feet to approximately 23 feet bgs at the basalt bedrock contact. In general, creosote-impacted soil decreases in vertical thickness and is encountered at deeper intervals within the native soil stratum as distance increases from the former wood-treating operations.

Creosote-related hydrocarbons (diesel and oil ranges) were detected in soil samples collected from 19 of 28 explorations completed in Area 1, with a maximum concentration of 6,320 mg/kg (sum of diesel- and oil-range hydrocarbons) in the MW-24 boring. Concentrations of total PAHs in soil ranged from not detected to a maximum of 3,410 milligrams per kilogram (mg/kg) in boring MW-24.

3.1.4 Sediment

The in-water areas were divided into multiple areas during the remedial investigations. However, this ROD focuses on the four in-water PAAs (Area 1 Dock, Area 2 Dock, Cove Area, and Upper Milton Creek). Since the 2000 remedial investigation report, over three hundred additional sediment samples from the Scappoose Bay and Milton Creek shorelines along Area 1 have been analyzed for PAHs, TPH, VPH/EPH, VOCs, TOC, and/or black carbon. Additionally, subsurface information was obtained via the 67 borings and depth discrete sampling, various sediment sampling events, and multiple shoreline inspections. Analytical results from sediments samples

collected during the supplemental remedial investigation are representative of current Site conditions and are summarized below. Sample locations are shown on Figures 32 and 33.

Petroleum hydrocarbons were detected in sediment samples collected along the Scappoose Bay shoreline. These were primarily diesel-range hydrocarbons with some oil-range hydrocarbon detections. A maximum of 14,000 mg/kg of diesel-range hydrocarbons was detected in sample SB-11 at 4 feet bml. Sample SB-11 was located within the estimated NAPL extent just west of the peninsula.

Total PAHs in the 2011 to 2013 surface sediment samples ranged from not detected to a maximum of approximately 4,800 mg/kg in SSED-6 located east of the former hog fuel dock within an area of creosote wood debris. Total PAHs in the 2017 surface sediment samples, all obtained outside inferred NAPL areas, ranged from 1.58 mg/kg (PWS-090517-6) to 53.54 mg/kg (PWS-090517- 9). As stated above, the highest relative total PAH concentrations were found within the old buried operational surface and/or within the inferred areas of offshore NAPL.

TOC concentrations in surface sediment samples ranged from 520 mg/kg at SB-23 to 150,000 mg/kg at SSED-6. These locations were east of the former hog fuel dock and within 25 feet of each other. The average TOC concentration was approximately 21,000 mg/kg. TOC content varied across the investigated area, and there does not appear to be a discernable pattern of high or low concentrations by locality. In general, samples with higher TOC were observed to contain wood debris, including wood chips, sticks, and roots.

The percentage of black carbon in surface sediment samples ranged from 0.10 to 0.34 (with the exception of one location at which black carbon was not detected), with an average of 0.18 percent (%) black carbon. Similar to TOC, black carbon content varied across the investigated area with no discernable pattern of high or low concentrations by locality.

As expected, the highest concentrations of PAHs in sediment occurred within the area of the inferred extent of NAPL, although NAPL was not observed in all samples collected within this area. Although TPH was not tested in sediment as frequently as PAHs, the highest TPH detections were also observed within the inferred extent of offshore NAPL.

3.1.5 Sediment Porewater

Sediment porewater samples were only collected during the supplemental remedial investigation. The majority of porewater samples were collected outside of the offshore inferred NAPL areas or outside of the in-water PAAs.

All sediment porewater concentrations reported from the ex-situ Solid-Phase Microextraction (SPME) and in-situ low density polyethylene (LDPE) passive samplers, are freely dissolved water concentrations (C_{free}) . In accordance with EPA guidelines, the direct measurements of the freely dissolved 34 PAH concentrations in sediment porewater or interstitial water (IW) are used along with their expected final chronic value (FCV) water/lipid partitioning behavior to calculate a hazard quotient, referred to as a toxic unit (TU). The narcosis based Tus are considered additive, and evaluation of the sum of the Tus requires that all 34 PAH analytes be included in the sum of IW TU (ΣIWTU). Calculations of ΣIWTU have been used on this project as a benchmark for

predicting the toxicity of PAHs to benthic invertebrates with a potential for unacceptable risk when the ΣIWTU is greater than or equal to (\ge) 1.

The results of depth-discrete porewater sampling were used to evaluate dissolved phased mass flux from deeper impacted sediment. In general, detected concentrations of dissolved PAHs in "deep" sediment porewater beneath the biologically active zone (i.e., 24 to 29 inches bml) were higher than those measured in the overlying "shallow" sediment and surface water. Porewater sampling conducted beyond the delineated areas of NAPL indicate that creosote constituents attenuate rapidly in the sediment pore space between 2 and 1 foot bml.

3.1.6 Surface Water

Surface water analytical results from 2017 are representative of current Site conditions and were used to determine nature and extent. Low levels of C_{free} PAHs were detected in all eight 2017 surface water samples. Total PAHs measured on the in-situ LDPE passive samples after 20 days in surface water ranged from 0.0192 μg/L (surface water sample co-located with PWS-090517-3) to 0.0797 μg/L (surface water sample co-located with PWS-090517-10).

The 18 parent PAH and 16 alkylated PAH homologue compounds (34 PAHs) results in surface water were compared to the FCVs and their corresponding Tus summed (ΣSWTU). Surface water samples co-located with PWS-090917-9, PWS-090517-10, PWS-090617-12, and PWS-090617- 13 produced a sum of surface water toxic unit (Σ SWTU) \geq 1.

In general, 34 PAH levels in the 2017 surface water samples were either consistent or slightly higher than detected concentrations in co-located shallow sediment porewater. The two notable exceptions to this data trend are PWS-090617-2 and PWS-090617-12 (located in Lower Milton Creek and the Area 1 Dock, respectively), where detected concentrations of PAHs in shallow porewater are higher than levels measured in the overlying surface water samples.

3.1.7 Air

Airborne dust transport is not a likely significant contaminant migration pathway. The original ground surface at the time the former wood-treating facility operated is currently located beneath 5 to 10 feet of non-contaminated dredge sand fill.

Gravel roads built atop the clean fill (at current ground surface) reportedly had oil applied to them between 1974 and 1991 for dust suppression. PCBs were detected in surface soil samples collected from former oiled gravel roadbeds at the Site.

3.2 RISK ASSESSEMENT

The standards for a protective cleanup are defined in the ORS and OAR. ORS 465.315 sets standards for degree of cleanup required, risk protocol, hot spots of contamination, etcetera (etc.). OAR 340-122-0084 describes the requirements for risk assessments while OAR 340-122-0115 provides additional definition of protectiveness.

Baseline risk assessments were initially conducted in 2006 to evaluate risks to human health and ecological receptors. The results of the baseline risk assessments are summarized in the Baseline Human Health Risk Assessment (Bridgewater Group and Kennedy/Jenks Consultants, 2006a) and the Baseline Ecological Risk Assessment Summary Report (Bridgewater Group and Kennedy/Jenks Consultants, 2006b). Updated supplemental HHRA and ERAs, incorporating new data as well as additional exposure scenarios, were completed in 2020 and are summarized in Appendices F and G of the Updated Supplemental Remedial Investigation Report, respectively (Cascadia Associates, 2020a; Cascadia Associates, 2020b).

The results of the risk assessments that have been conducted for human health and ecological receptors at the former Pope & Talbot site are summarized below. The residual risk associated with the selected remedy is summarized in Section 10.2.

3.2.1 Conceptual Site Model

This ROD addresses contamination in Area 1 Upland as well as within the nearshore sediment areas adjacent to the Site within Scappoose Bay and Milton Creek. The primary source of contamination at the Site is releases from historical industrial activities during plant operation (see Section 2.3). As discussed in Section 3.1, creosote is the primary contaminant in soil, groundwater, and sediment samples collected in the vicinity of the former wood-treating facilities. During plant operations, leaks and spills of creosote NAPL seeped into the ground and collected in the native soil above the basalt bedrock within an approximately 4-acre portion of the Area 1 Upland. A schematic of the CSM is shown on Figure 39.

Creosote is a multi-component NAPL that contains many hydrocarbons, primarily PAHs, phenolic compounds, and carrier fluids such as diesel. In the approximately 60 years since the plant closure, subsurface creosote (either as free phase NAPL, sorbed onto solids, or dissolved in water) appears to have undergone compositional changes due to weathering (e.g., loss of more soluble or biodegradable components). Subsequently, residual Site contamination consists predominantly of creosote-derived PAHs (i.e., no light NAPL or evidence of carrier fluids).

Another significant source of in-water impacts is residual creosote-contaminated wood debris from historical operations and historical infrastructure (timber piling and dock structures). Much of this wood debris appears to have been shed from the upland, over-water operations, and/or from log rafts historically moored along the shoreline. The woody material, much of it saturated with creosote, settled out over time and eventually was incorporated into the offshore sediment profile. In general, Site use and tenant activities since the wood treating operations ceased in 1959 do not appear to have significantly caused or contributed to the contamination levels encountered during recent investigations.

Current contamination sources in the upland include contaminated groundwater and soil. Volatile compounds in groundwater have the potential to volatilize into indoor and outdoor air. Contaminated groundwater has the potential to migrate to surface water, sediment, sediment porewater, and surface water in the in-water area. Current contamination sources in the in-water area include sediment/beach soil and contaminated woody debris.

The **CSM for evaluation of human health risks** is summarized in Figure 40. The following human health exposure scenarios are considered potentially complete for the Site:

- Industrial workers through ingesting and coming into direct contact with upland surface soil; inhaling contaminants volatilized from upland surface and subsurface soil; ingesting upland surface soil; and inhaling contaminants volatilized from upland groundwater.
- Excavation workers through ingesting and coming into direct contact with upland surface and subsurface soil; inhaling contaminants volatilized from upland surface and subsurface soil; ingesting upland surface and subsurface soil; and direct contact with upland groundwater.
- Construction workers through ingesting and coming into direct contact with upland surface and subsurface soil; inhaling contaminants volatilized from upland surface and subsurface soil; ingesting upland surface and subsurface soil; and direct contact with upland groundwater.
- Sport fishers through ingesting fish and shellfish impacted by contaminants bioaccumulated from surface water; ingesting and coming into direct contact with beach soil/sediment; inhaling contaminants volatilized from beach soil/sediment; and ingesting surface water.
- Subsistence fishers through ingesting fish and shellfish impacted by contaminants bioaccumulated from surface water; ingesting and coming into direct contact with beach soil/sediment; inhaling contaminants volatilized from beach soil/sediment; and ingesting surface water.
- Recreational trespassers, transients, and adult boaters through ingesting and coming into direct contact with beach soil/sediment; inhaling contaminants volatilized from beach soil/sediment; coming into direct contact with surface water; ingesting surface water; and coming into direct contact with sediment/sediment porewater.

The **CSM for evaluation of ecological risks** is summarized in Figure 41.

The following **terrestrial** ecological exposure scenarios are considered complete or potentially complete for the Site:

- Upland plants through direct contact with upland surface and subsurface soil.
- Soil invertebrates through ingestion of and direct contact with upland surface and subsurface soil.
- Herbivorous birds and mammals through ingestion of and direct contact with upland surface and subsurface soil; ingestion of upland biota; ingestion of and direct contact with beach soil/near-shore sediment; and ingestion of and direct contact with surface water.
- Insectivorous birds and mammals through ingestion of and direct contact with upland surface and subsurface soil; ingestion of upland biota; ingestion of and direct contact with beach soil/near-shore sediment; and ingestion of and direct contact with surface water.
- Carnivorous birds and mammals through ingestion of and direct contact with upland surface and subsurface soil; ingestion of upland biota; ingestion of and direct contact with beach soil/near-shore sediment; and ingestion of and direct contact with surface water.

The following **aquatic** ecological exposure scenarios are considered complete or potentially complete for the Site:

- Wetland and aquatic plants through direct contact with beach soil/near-shore sediment, offshore sediment, sediment porewater, and surface water.
- Benthic invertebrates through ingestion of and direct contact with beach soil/near-shore sediment, offshore sediment, and surface water; direct contact with sediment porewater; and ingestion of aquatic biota.
- Fish, through ingestion of and direct contact with beach soil/near-shore sediment, offshore sediment, and surface water; direct contact with sediment porewater; and ingestion of aquatic biota.
- Insectivorous birds, through ingestion of and direct contact with upland surface soil, beach soil/near-shore sediment, and surface water; and ingestion of upland and aquatic biota.
- Piscivorous mammals and birds through ingestion of and direct contact with upland surface and subsurface soil; ingestion of upland biota; ingestion of and direct contact with beach soil/near-shore sediment, ingestion of and direct contact with surface water; and ingestion of aquatic biota.

3.2.2 Human Health Risk Screening

The HHRA evaluated PAHs, TPH, select VOCs, select metals, PCBs in soil, and dioxins and furans in sediment, as requested by DEQ (Cascadia Associates, 2020a). The constituent classes evaluated in the human health risk screening for each environmental medium are summarized below:

Air. No air data have been evaluated from the Site.

Sediment inside the area of inferred NAPL. The Updated Supplemental HHRA does not evaluate data from sediment samples collected within areas of inferred NAPL because exposure to NAPL is considered to have unacceptable risk and therefore does not require further evaluation.

Sediment outside the area of inferred NAPL. PAHs, TPH, and dioxins and furans were evaluated. The 2006 HHRA evaluated discrete and composite sample results collected in 1996, as well as a small number of samples collected in 2005, but it excluded multiple
sediment samples collected in 2003 and 2004. Therefore, DEQ determined that a comprehensive quantitative assessment of all available sediment data was appropriate in the supplemental HHRA (DEQ, 2014a).

Soil. To adequately reflect current risk assessment guidance, risk from PCBs in upland soil was re-evaluated in the supplemental HHRA using current risk-based criteria (DEQ, 2010) for all Site receptors (industrial worker, excavation worker, and construction worker). Per DEQ's request, risk from select COPCs to construction workers was also reassessed.

Sediment Porewater. Human receptors are not exposed to sediment porewater as a separate media and a separate evaluation of porewater was not conducted (DEQ, 2014b), except to compare porewater concentrations to screening levels values (SLVs) protective of people consuming fish that have accumulated Site-related chemicals.

Surface Water. Eight surface water samples were collected concurrently with porewater samples in 2017 and analyzed for 34 PAHs. These concentrations were screened against Preliminary Remediation Goals (PRGs), as shown in Table F-2E of the Supplemental HHRA (Cascadia Associates, 2020a). Surface water samples were also evaluated for protection of people consuming fish from carcinogenic PAHs by comparing them to the same Site-specific SLVs as for porewater.

Groundwater. The 2006 HHRA evaluated groundwater data collected between 1996 and 2004 and concluded that the potential exposure scenario of groundwater exposure by an excavation worker resulted in risk estimates exceeding DEQ acceptable risk levels. PAHs were therefore retained as a COC in groundwater based on the findings of the 2006 HHRA (Bridgewater Group and Kennedy/Jenks Consultants, 2006a). The Updated Supplemental HHRA evaluated all groundwater data collected from Area 1 Upland between 2005 and 2012, including those collected from monitoring wells within the inferred extent of subsurface NAPL. Monitoring well locations and inferred extent of subsurface NAPL are shown on Figure 39. Groundwater from Area 2 Upland was not evaluated because this portion of the Site received an NFA determination from DEQ in 2008.

The 2020 Supplemental HHRA evaluated risks associated with PAHs, TPH, BTEX, and select metals (arsenic, cadmium, chromium, lead, and zinc) in groundwater.

3.2.3 Human Health Risk Assessment

The 2006 Baseline HHRA and 2020 Supplemental HHRA (Bridgewater Group and Kennedy/Jenks Consultants, 2006a; Cascadia Associates, 2020a) describe in detail the procedures used to evaluate the potential risks to human health associated with the chemicals in the media described above. The findings of the HHRAs are summarized below.

Chemicals of Potential Concern.

The HHRAs identified the following COPCs for each of the contaminated media on the Site:

Sediment. All PAHs, TPH, and dioxins and furans detected in sediment were retained as sediment COPCs.

Soil. COPCs from the 2006 HHRA were retained as COPCs in the Updated HHRA (acenaphthene, methylene chloride, benzo(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene, chrysene, indeno(1,2,3-cd)pyrene, 2 methylnaphthalene, fluoranthene, fluorene, naphthalene, phenanthrene, pyrene, PCP, arsenic, and lead). A full list of selection criteria for COPCs is provided in Table 1 of the Baseline HHRA (Bridgewater Group and Kennedy/Jenks Consultants, 2006a). PCBs were also retained as COPCs in soil in the Supplemental HHRA.

Surface Water. PAHs were retained as surface water COPCs.

Groundwater. The selection of groundwater COPCs from the supplemental HHRA is illustrated in Table F-1 of the Supplemental HHRA (Cascadia Associates, 2020a). Eleven PAHs and TPH were retained as groundwater COPCs. No metals or VOCs were retained as groundwater COPCs.

Pathway Analysis. The Baseline HHRA identified the following primary human health exposure pathway of concern: potential future on-site excavation workers exposed to PAHs in contaminated groundwater in Area 1 Upland of the Site. The Supplemental HHRA further evaluated exposures to recreational and subsistence anglers consuming fish from Scappoose Bay, in addition to the rest of the exposure scenarios listed in Section 3.2.1, incorporating more recent data and subsequent input from DEQ.

Cumulative Risk. The Supplemental HHRA evaluated cancer risk using the Central Tendency Exposure (CTE) and Reasonable Maximum Exposure (RME) methods to evaluate risks to human health associated with the Site for on-site industrial workers, on-site excavation workers, on-site construction workers, recreational trespassers in Scappoose Bay, recreational trespasser in Milton Creek, sport and subsistence fishers in Scappoose Bay, and sport and subsistence fishers in Milton Creek (see Table 2).

3.2.3.1 Cumulative Human Health Risk Estimates

To evaluate cumulative cancer risk, the Total Lifetime Excess Cancer Risk was calculated for each receptor. To evaluate non-cancer risk, the total hazard indices were calculated for each receptor. Resulting cumulative cancer and non-cancer risk estimates for each receptor are discussed in detail in Section 6 of the Supplemental HHRA and are presented in Table 2.

Excess lifetime cancer risk estimates did not exceed Oregon's environmental cleanup standard of 10^{-6} for exposure to individual carcinogenic substances or 10^{-5} for multiple carcinogenic substances for the following human health exposure scenarios: industrial workers, excavation workers, construction workers, recreational trespassers in Scappoose Bay, recreational trespassers in Milton Creek, or sport fishers in Scappoose Bay. Excess cancer risk estimates were not derived for sport or subsistence fishers in Milton Creek due to insufficient sediment data.

Non-cancer hazard indices did not exceed Oregon's environmental cleanup standard of one for the following exposure scenarios: recreational trespassers in Scappoose Bay, recreational trespassers in Milton Creek, sport fishers in Scappoose Bay, subsistence fishers in Scappoose Bay, sport fishers in Milton Creek, or subsistence fishers in Milton Creek.

The following section summarizes risk estimates exceeding Oregon's environmental cleanup standard for exposure to carcinogenic and non-carcinogenic substances for each receptor exposure scenario.

Cumulative Cancer Risk Estimate Exceedances.

Adult/child subsistence fishers in Scappoose Bay. For adult/child subsistence fishers in Scappoose Bay, the excess cancer risk estimate exceeded Oregon's environmental cleanup standard of 10^{-6} for exposure to individual carcinogenic substances or 10^{-5} for multiple carcinogenic substances. The total excess cancer risk estimate was $6x10^{-5}$ and $3x10^{-5}$ for RME and CTE (see Table 2). The HHRA recognized that both recreational and subsistence anglers could provide fish to children so both adult and child receptor scenarios were considered. These risk estimates are based on bioaccumulation of COPCs from sediment, surface water, and/or porewater. There is unacceptable risk for the subsistence fisher from the consumption of fish exposed to Site-related dioxins and furans in the Scappoose Bay area; exceedances were mostly limited to detected concentrations in samples collected from within the extent of known petroleum sheen and are therefore anticipated to be co-located with hydrocarbons in the areas of petroleum sheen (Cascadia Associates, 2020a, Section 6.6; Table F-2C). Exposure scenarios for bioaccumulation from surface water and sediment indicate a low level of concern for bioaccumulation of carcinogenic PAHs. Bioaccumulation from porewater is considered a secondary line of evidence (LOE) and was used as a supplement to evaluate risk.

Cumulative Non-Cancer Risk Estimate Exceedances.

Industrial workers. For industrial workers, the total hazard index exceeded Oregon's environmental cleanup standard of one. The total hazard indices were 770 and 54 for RME and CTE, respectively (see Table 2).

Elevated risk estimates are dominantly associated with industrial workers inhaling contaminants volatilizing from groundwater into indoor and outdoor air in the upland. The risks associated with this pathway are based on exposure to volatilization of petroleum hydrocarbon concentrations in localized groundwater samples (MW-3A and MW-22) within the extent of inferred subsurface NAPL in Area 1 Upland (Cascadia Associates, 2020a, Tables F-6A, F-6B, F-6C, F-6D). The inhalation risk associated with volatilization to outdoor and indoor air contributes nearly 100% of the potential risk to the total non-cancer risk estimates for industrial workers.

Direct contact with surface soil also poses potential risk to industrial workers. This risk is based on exposure to PCBs in a limited number of soil samples collected in the upland (SS-08, SS-10, SS-19, SS-22, and SS-24). The direct contact with soil pathway contributes relatively little to the cumulative estimated non-cancer risk to industrial workers. Specifically, hazard indices for this pathway were 2 and 0.5 for RME and CTE, respectively (Cascadia Associates, 2020a, Tables F-13E and F-13F).

Excavation workers. For excavation workers, the total hazard index exceeded Oregon's environmental cleanup standard of one. The total hazard indices were 390 and 28 for RME and CTE, respectively (see Table 2).

Estimated non-cancer risks to the excavation worker are dominated by direct contact with and inhalation of petroleum hydrocarbons in groundwater (Cascadia Associates, 2020a, Tables F-7A, F-7B). Direct contact with surface soil, and direct contact, ingestion, and inhalation of subsurface soil also contribute to the cumulative estimated risk to potential excavation workers (Cascadia Associates, 2020a, Tables F-13C, F-13D).

No unacceptable risks were estimated for direct exposure to soil, or for indirect exposure to contaminants in soil from inhalation outside the inferred NAPL areas, with the exception of an exceedance of risk calculated due to the localized high concentration of arsenic in one deep soil sample (MW-10). The potential risks from direct exposure to groundwater in an excavation were not quantified because the COPCs are semi-volatile. DEQ does not recommend a quantitative assessment of dermal exposures to SVOCs because their dermal permeability coefficients are outside of the effective predictive domain used to model this exposure factor. Instead, a qualitative discussion was provided in the uncertainty section of the Supplemental HHRA and concluded that:

- It is reasonable to assume there are unacceptable risks or hazards from dermal contact with groundwater where NAPL is present.
- It is not possible to accurately quantify the magnitude of the risk from dermal contact with water containing appreciable levels of PAHs or other semi-volatile compounds (such as TPH consisting of weathered creosote) in areas where NAPL is not present.
- Thus, an administrative or engineering control should be applied to the Site to control future exposures to groundwater. Administrative and engineering controls have already been established for Area 2 and similar controls could easily be applied to Area 1 of the Site.

Construction workers. For construction workers, the total hazard index exceeded Oregon's environmental cleanup standard of one. The total hazard indices were 17 and 0.6 for RME and CTE, respectively (see Table 2).

Estimated non-cancer risks to the construction worker are dominated by direct contact, ingestion, and inhalation of subsurface soil (Cascadia Associates, 2020a, Tables F-13A, F-13B). Direct contact with surface soil also contributes to the cumulative estimated risk to potential excavation workers (Cascadia Associates, 2020a, Tables F-13A, F-13B). This risk estimate is based on exposure to the localized high concentration of arsenic in one deep soil sample (MW-10). With the exception of this one sample from MW-10, no unacceptable risks were estimated for a potential construction worker's direct exposure to soil, or from indirect exposure to contaminants in soil, outside the inferred NAPL areas.

3.2.4 Ecological Risk Assessment

The ERA reports (Bridgewater Group and Kennedy/Jenks Consultants, 2006b; Cascadia Associates, 2020b) describe in detail the procedures used to evaluate the potential risks to ecological receptors at the Site. A Level II ERA was conducted to evaluate terrestrial and aquatic ecological exposures associated with Site-related contamination (Bridgewater Group and Hart Crowser, 2003). The 2003 Level II ERA identified contaminants of potential ecological concern (CEPCs) and appropriate ecological assessment endpoints. The Level II ERA concluded that no CPECs are present in surface soil and therefore that no further ERA activities for terrestrial receptors are necessary at the Site. The Level II ERA identified potential unacceptable risk to ecological receptors in the in-water areas of Scappoose Bay and Milton Creek adjacent to the Site (Bridgewater Group and Hart Crowser, 2003; Bridgewater Group and Kennedy/Jenks Consultants, 2006b, Section 2).

Based on the conclusions of the Level II ERA, the 2006 ERA Report evaluated risks to aquatic ecological receptors associated with contaminant data for Scappoose Bay sediment, surface water, and seeps. This assessment concluded that contaminant concentrations in Scappoose Bay seeps may pose unacceptable risk to ecological receptors and identified that preventing ecological receptors from direct contact with Scappoose Bay seeps should be a preliminary remedial action goal for the Site (Bridgewater Group and Kennedy/Jenks Consultants, 2006b, Section 3).

At DEQ's request, the 2020 Supplemental ERA re-evaluated the CSM for the entire site; evaluated ecological risks using sediment, porewater, and surface water data collected since 2005; and applied new methods and research to the data evaluation as additional LOEs supporting ecological risk conclusions. The presence of NAPL is presumed to present an unacceptable risk for ecological receptors. Therefore, ecological risks associated with media within the estimated extent of NAPL were not quantitatively evaluated in the Supplemental ERA (Cascadia Associates, 2020b).

3.2.4.1 Ecological Risk Assessment Lines of Evidence

The Supplemental ERA used multiple LOEs to evaluate ecological risk. The first and primary LOE was the presence or absence of a surface sheen. Additional LOEs included comparisons of environmental concentrations in various environmental media to SLVs based on toxic responses of aquatic organisms (e.g., benthic organisms and fish) to individual chemical constituents. The LOEs used in the Supplemental ERA are discussed below:

Presence or absence of sheen: The presence of NAPL and/or a petroleum sheen was used as a primary LOE in evaluating risk to aquatic receptors. The presence of NAPL or a sheen is a violation of narrative water quality criteria and is considered an unacceptable risk to ecological receptors due to the potential for fouling of membranes and gills or ingestion of accumulated product on feathers or fur during preening. Multiple field events (2010, 2011, 2012, 2013, 2017, and 2018) were conducted to properly delineate the boundary of NAPL/sheen, and the results are shown on Figure 42.

Sediment toxicity:

Total PAHs: Bulk sediment concentrations of total PAHs were compared with the threshold effects concentration (TEC) and probable effects concentration (PEC) to evaluate risk to benthic organisms associated with direct contact with sediment (Macdonald et al., 2000). TECs and PECs were developed from a number of published sediment quality guidelines (SQGs). TECs reflect lower limits of sediment toxicity where toxicity is not predicted to occur; PECs reflect upper limits, above which adverse effects are predicted to have a high probability of occurring. Evaluating bulk sediment concentrations against both TECs and PECs provides a bracket for expected toxicity and provides a secondary LOE in assessing risk of sediment toxicity to the benthos. Concentrations of total PAHs and a comparison to PEC and TEC values are presented in Table G-10 of the Supplemental ERA (Cascadia Associates, 2020a).

Total petroleum hydrocarbons: PECs and TECs do not exist for TPH. TPH fractions were screened against criteria developed as part of the Portland Harbor Superfund Site's (PHSS's) ERA (Windward, 2013), with values updated to correct a calculation error. Use of PHSS's SLVs are considered a secondary LOE in risk evaluation. Concentrations and SLVs are presented in Table G-10 of the Supplemental ERA (Cascadia Associates, 2020a).

Porewater toxicity:

PAHs: Concentrations of total PAHs in porewater samples were converted into total TU values, specifically the ΣIWTU, to evaluate the potential for narcotic risk to benthic invertebrates. Each individual PAH congener concentration was divided by its unique final chronic value (based on narcotic effects to benthic organisms) to calculate the TU. The sum of all Tus for a sample was calculated as the ΣIWTU. ΣIWTU values ≥1 indicate potential risk to benthic organisms. ΣIWTU values are presented in Table G-6a and Table G-11a of the Supplemental ERA (Cascadia Associates, 2020a).

VPH: Concentrations of VPH in 2017 porewater samples were used as a secondary LOE for determining risk at the Site. Concentrations were compared to PHSS's ERA SLVs and presented in Table G-11b of the Supplemental ERA (Cascadia Associates, 2020a).

OPAHs: OPAHs were measured in 2017 porewater samples to provide a secondary LOE for determining risk at the Site. The presence of oxygen atoms in OPAH molecules makes OPAHs more polar and hydrophilic than PAHs, often resulting in higher freely dissolved concentrations of OPAHs than PAHs in sediment porewater. No SLVs exist for OPAHs, so there was no quantitative assessment of risk based on OPAH concentrations. Concentrations of OPAHs in surface water samples are presented in Table G-12a of the Supplemental ERA (Cascadia Associates, 2020a).

Surface water toxicity:

PAHs: As with porewater, concentrations of PAHs in surface water samples were converted to Tus. The sum of all Tus for a given sample were calculated as the sum of surface water TU (Σ SWTU). A Σ SWTU greater than 1 indicates the potential risk to benthic organisms, and these calculations were used as a primary LOE in risk evaluation. ΣSWTUs are presented in Table G-6b and G-12a of the Supplemental ERA (Cascadia Associates, 2020a). Concentrations of C_{free} PAHs in surface water were also compared to published SLVs for individual PAHs where available and presented in Table G-12b of the Supplemental ERA (Cascadia Associates, 2020a).

OPAHs: As with porewater, OPAHs were measured in a number of samples to further define risk from PAHs. Concentrations of OPAHs in surface water were not screened against any SLVs, as no risk-based screening criteria exists. Total OPAH concentrations are presented in Table G-12a of the Supplemental ERA (Cascadia Associates, 2020a).

Groundwater toxicity: Groundwater was evaluated to determine if upland dissolved phase constituents have the potential to adversely affect Scappoose Bay. Dissolved phase concentrations in groundwater were evaluated based on the same approach used for surface water and porewater. This approach was completed for nearshore groundwater wells and did not account for any potential change in concentrations during migration to Scappoose Bay.

PAHs: As with porewater and surface water, concentrations of PAHs in groundwater were converted to Tus. The sum of all Tus in a given sample were calculated as the ΣGWTU. A ΣGWTU greater than 1 indicates the potential risk to benthic organisms. As mentioned above, benthic organisms are not directly exposed to groundwater; thus, this analysis is strictly used to examine the potential for risk to benthic organisms following upland groundwater migration to Scappoose Bay. PAH concentrations and ΣGWTUs are presented in Table G-13 of the Supplemental ERA (Cascadia Associates, 2020a). Concentrations of individual PAHs in groundwater were compared to DEQ Level II SLVs (DEQ, 2001b) to evaluate potential risk after upland groundwater migration to Scappoose Bay. Concentrations and associated SLVs are presented in Table G-13 of the Supplemental ERA (Cascadia Associates, 2020a).

TPH: Concentrations of diesel- and gasoline-range hydrocarbons were compared to Freshwater Ecotox Aquatic Habitat Goal Levels from San Francisco Bay Regional Water Quality Control Board (SFBRWQCB, 2019). As described above, this approach was used as a conservative analysis of potential risk to Scappoose Bay from upland groundwater migration. Concentrations and comparisons to SLVs are presented in Table G-13 of the Supplemental ERA (Cascadia Associates, 2020a).

3.2.4.2 Ecological Risk Assessment Conclusions

The following conclusions were drawn from the Supplemental ERA:

- The presence of creosote NAPL and moderate to heavy petroleum sheen present unacceptable risk to ecological receptors where they are observed in surface sediment, porewater, and surface water. Areas exhibiting moderate to heavy petroleum sheen in surface sediment are shown on Figure 42.
- Issues with the 2012 measurement of PAHs C_{free} in sediment porewater centered around elevated detection limits, particularly high molecular weight PAHs, and a perceived low bias for detecting select low molecular weight PAHs (e.g., naphthalene and its alkylated isomers). This uncertainty to confidently assess porewater toxicity led DEQ to consider other LOEs. Subsequently, passive samplers were deployed in 2017 within offshore areas where prior efforts to assess porewater toxicity and habitat impairment posed by Site -related contamination had been inconclusive. The quantitation of chemicals in

sediment porewater using LDPE passive sampling devices developed by the Food Safety and Environmental Stewardship program at Oregon State University (OSU) is more robust and precise than the 2012 direct measurements of freely dissolved PAHs in sediment porewater. For example, naphthalene, a key component of PAH contamination in upland groundwater, was detected in 26 of 30 porewater samples collected using passive samplers in 2017, compared to only 4 out of 59 porewater samples analyzed by the ASTM International SPME approach in 2012.

- The 2017 porewater data demonstrate low or nonexistent potential narcotic risk to benthic organisms outside of delineated offshore NAPL areas. Specifically, the ΣIWTU exceeded in only one of the 15 Phase 3 locations, PWS-090617-2, based on porewater measurements obtained within the biologically active zone (i.e., upper 12 inches of sediment). Sediment porewater sample PWS-090617-2 was obtained in Lower Milton Creek near the confluence with Scappoose Bay. This area is not currently included in the in-water PAAs and therefore may need to be addressed via additional sampling and action moving forward.
- Like porewater, surface water samples with a Σ SWTU ≥ 1 are considered unacceptable. Four of the eight 2017 surface water passive sampling sites (PWS-090917-9, PWS-090517- 10, PWS-090617-12, and PWS-090617-13) resulted in a ΣSWTU slightly above 1 (1.07). Detected concentrations of alkyl PAHs C4-naphthalene appear to be the primary ecological risk driver in surface water.
- Risk to fish is negligible for the Scappoose Bay exposure area from either direct exposure to surface water or from the bioaccumulation of fluoranthene or pyrene.
- A screening of monitoring wells along Scappoose Bay indicates dissolved phase constituents may pose a risk to aquatic receptors from discharge of upland groundwater to surface water based on exceedances of select SLVs for PAHs and TPH and ΣGWTU values greater than 1. There is still a large degree of uncertainty associated with this pathway; further evaluation is needed, including additional sampling at the Site of groundwater discharge.

Taken together, along with the results of earlier ERAs, bioassay results, and the multiple lines of evidence considered over more than two decades of investigation, the 2020 ERA demonstrates:

- The presence of NAPL and/or petroleum sheen present unacceptable risk for ecological receptors.
- Dissolved phase groundwater constituents may pose risk to aquatic receptors from discharge of upland groundwater to the aquatic environment; consistent with the findings presented in the 2006 ERA, unacceptable risk to ecological receptors from direct contact with groundwater migrating to Scappoose Bay via seeps should be a preliminary remedial action goal for the Site.
- There is low or nonexistent potential risk to ecological receptors at the Site outside of delineated areas of NAPL and petroleum sheen.

3.2.5 Human Health and Ecological Risk Summary and RAOs to Address Risk

This section connects the unacceptable risks identified in the HHRAs and ERAs with RAOs developed to focus the remediation on addressing those risks. In this section the RAOs are referenced in context of the risks they aim to mitigate; the RAOs are described in more detail in Section 5.1.2.

3.2.5.1 Human Health Risk Summary

The HHRA identified unacceptable risks associated with Area 1 Upland related contamination in the following human health exposure scenarios:

- Adult/child subsistence fishers in Scappoose Bay.
- Industrial workers working in the upland area of the Site.
- Excavation workers working in the upland area of the Site.
- Construction workers working in the upland area of the Site.

3.2.5.2 RAOs to Address Human Health Risk

Below is a summary of RAOs selected for the remedy to address unacceptable risks to human health:

Adult/child subsistence fishers in Scappoose Bay.

A core focus of the remediation is on limiting bioaccumulation of COPCs in sediment that drive the excess lifetime cancer risk estimate for subsistence fishers, and on limiting bioaccumulation of carcinogenic PAHs in surface water and sediment that contribute risk to subsistence fishers. **RAO 1a, RAO 1c, and RAO 2** all directly address this goal by: preventing releases of creosote NAPL (defined as moderate to heavy petroleum sheen) to the aquatic environment; preventing unacceptable risk to aquatic receptors associated with direct contact with NAPL in riverbank seeps, surface sediment, and surface water; and protecting aquatic receptors from exposure to contaminants in surface sediment (as defined as the top 12 inches of sediment and associated porewater) and surface water, respectively. **RAO 4** also addresses risk associated with in-water receptors bioaccumulating contaminants migrating from the groundwater in the upland to the aquatic environment by preventing recontamination of the in-water remedy from the defined NAPL riverbank seep areas and contaminated groundwater discharge. **RAO 1b** also addresses potential risk to adult/child fishers and other recreational users by preventing unacceptable risk to recreational users associated with direct contact to NAPL seeps in surface sediment. The selected remedial alternative for the Area 1 Upland, Area 1 Dock PAA, Area 2 Dock PAA, and Cove Area PAA address this risk by limiting contaminants migrating from the upland into the in-water areas via seeps and soil erosion, and through a combination of removing and capping hot spots of contamination in the inwater PAAs.

Industrial workers working in the upland area of the Site.

RAO 8 directly addresses risk to industrial workers by protecting Site workers from vapor intrusion into buildings that causes unacceptable risk. The selected remedy for Area 1 Upland will cap upland soil contamination and address contaminant migration into the aquatic environment. The remedy does not, however, actively remove contamination in upland groundwater or soil in the source area. The long-term effectiveness of the selected Area 1 Upland remedy is therefore dependent on institutional and engineering controls to mitigate risks to potential industrial workers in the upland. The selected remedy for the upland assumes that an EES would be made between the Port of Columbia County and DEQ. The EES would include requirements to follow a contaminated media management plan (CMMP), require installation of engineering controls to prevent vapor intrusion into any future structures, land use restrictions, water use restrictions, identification and maintenance of engineered barriers, and long-term monitoring requirements.

RAO 6 directly addresses risk to industrial workers in areas of unacceptable risk related to PCB contamination in soil. The selected upland remedy will cap surface soil contamination in the Upland PAA, mitigating this risk to potential industrial workers.

Excavation workers working in the upland area of the Site.

RAOs 5, 6, and 7 directly address risk to excavation workers by protecting construction/excavation workers when working in upland area(s) of inferred NAPL and/or near monitoring well boring MW-10; preventing industrial workers in areas of unacceptable risk related to PCB contamination in gravel roadways; and preventing exposure to groundwater in Area 1 Upland by Site workers, respectively. As described above, the selected remedy for Area 1 Upland will cap soil contamination and address contaminant migration into the aquatic environment but will not actively remove contamination in upland groundwater or subsurface soil in the source areas. The longterm effectiveness of the selected remedial action is therefore dependent on institutional and engineering controls to mitigate risks to potential excavation workers in the upland. To be protective of a potential future excavation worker, this EES would need to mitigate or prevent: direct contact with and inhalation of contaminants from upland groundwater; and direct contact, ingestion, and inhalation of subsurface soil.

Construction workers working in the upland area of the Site.

RAO 5 directly addresses potential risk to construction workers by protecting construction/excavation workers when working in upland area(s) of inferred NAPL and/or near monitoring well boring MW-10. As described above, the selected remedy for the Area 1 Upland will cap upland soil contamination and address contaminant migration into the aquatic environment but will not actively remove contamination in upland subsurface soil. The long-term effectiveness of the Area 1 Upland selected remedial action is therefore dependent on institutional and engineering controls to mitigate risks to potential construction workers in the upland. To be protective of a potential future construction worker, this EES would need to mitigate or prevent direct contact, ingestion, and inhalation of subsurface soil.

3.2.5.3 Ecological Risk Summary

Based on the conclusions of the 2020 Supplemental ERA and an evaluation of all available LOEs, the following Site conditions were organized into primary and secondary indicators of aquatic ecological risk in the Feasibility Study (GeoEngineers, 2022):

Primary LOEs:

- Observations of a moderate to heavy petroleum sheen (i.e., creosote NAPL) in surface sediment (1 foot bml).
- Surface water concentrations representing a Σ SWTU ≥ 1 based on the direct measurement of 34 PAHs.
- Surface sediment porewater concentrations representing a $\Sigma IWTU \geq 1$ based on the direct measurement of 34 PAHs.

Secondary LOEs

- Observations of a slight petroleum sheen in surface sediment and subsurface sediment porewater (greater than [>]1 foot bml) concentrations representing a $\Sigma IWTU \geq 1$.
- Bulk sediment total 34 PAH concentrations \geq PEC of 22.8 mg/kg.
- Significant concentrations of total TPH in bulk sediment relative to fraction-specific SLVs.
- Bulk sediment total 34 PAH concentrations >10 mg/kg, representing the mid-point between the TEC of 1.6 mg/kg and the PEC of 22.8 mg/kg.
- Total TPH porewater concentrations greater than fraction-specific SLVs
- Porewater concentrations in surface sediment representing one-half of a TU.
- Relative magnitude of C_{free} 63 PAHs $> 0.5 \mu g/L$ (not risk-based).
- Relative magnitude of C_{free} total OPAHs concentrations $> 0.5 \mu g/L$ (not risk-based).

Based on these primary and secondary LOEs, the following conclusions were drawn in the Feasibility Study concerning ecological risk (GeoEngineers, 2022):

• Presence of creosote NAPL and moderate to heavy petroleum sheen presents unacceptable risk to ecological receptors where it is observed in surface sediment and surface water.

- Risk to ecological receptors in sediment porewater and surface water are largely driven by lighter end and more soluble PAHs, including naphthalene and methylnaphthalenes, C-3 and C-4 naphthalenes, C-2 and C-3 fluorenes, and C-2, C-3 and C-4 phenanthrenes/anthracenes. This conclusion is consistent for multiple pathways, including direct toxicity (e.g., comparison with Tier II final chronic values), narcosis (i.e., measured by toxic units), and bioaccumulation (i.e., using food chain multipliers) (DEQ, 2020).
- The presence of slight petroleum sheens in surface sediment and $\Sigma W T U \ge 1$ in subsurface sediment has the potential to result in unacceptable risk due to transformation of PAHs to more biologically active metabolites, and variability in the depth of the biologically active zone (DEQ, 2020).
- Total dissolved parent and alkylated PAHs in groundwater should be used to determine levels discharging to offshore sediment and surface water via colloidal groundwater transport, in addition to freely dissolved concentrations and NAPL migration.

3.2.5.4 RAOs to Address Ecological Risk

Based on the findings of the ERAs, the primary focus of the in-water remedy is on limiting aquatic ecological receptors' exposure to NAPL and petroleum sheen. **RAO 1a** directly addresses estimated ecological risk by preventing releases of creosote NAPL, defined as moderate to heavy petroleum sheen, to the aquatic environment. **RAO 1c** directly addresses this risk by preventing unacceptable risk to aquatic receptors associated with direct contact with NAPL in riverbank seeps, surface sediment, and surface water. **RAO 2** addresses this risk by protecting aquatic receptors from exposure to contaminants in surface sediment (as defined as the top 12 inches of sediment and associated porewater) and surface water that result in toxic effects, respectively. **RAO 3** addresses this risk by preventing recontamination of the in-water remedy from adjacent contaminated riverbank soils, further ensuring long-term protection of aquatic ecological receptors. **RAO 4** also addresses risk to in-water receptors associated with migration of contaminants from the groundwater in the upland by preventing recontamination of the in-water remedy from the defined NAPL riverbank seep areas and contaminated groundwater discharge.

3.3 BENEFICIAL USE AND HOT SPOT DETERMINATION

The criteria used to evaluate Remedial Action Alternatives (RAAs) for groundwater and surface water depend on whether a "hot spot" is present or not, as determined by a loss of "current or reasonably likely future" beneficial use of the water resource.

OAR 3401-122-0115(9) defines beneficial uses of water as any current or reasonably likely future beneficial use of groundwater or surface water by humans or ecological receptors.

OAR 340-122-0115(32) defines hot spot of contamination as:

(a) For groundwater or surface water, hazardous substances having a significant adverse effect on beneficial uses of water or waters to which the hazardous substances would be reasonably likely to migrate and for which treatment is reasonably likely to restore or protect such beneficial uses within a reasonable time, as determined in the feasibility

study; and (b) For media other than groundwater or surface water, (e.g., contaminated soil, debris, sediments, and sludges; drummed wastes; "pools" of dense, non-aqueous phase liquids submerged beneath groundwater or in fractured bedrock; and non-aqueous phase liquids floating on groundwater), if hazardous substances present a risk to human health or the environment exceeding the acceptable risk level, the extent to which the hazardous substances:

(A) Are present in concentrations exceeding: (i) 100 times the acceptable risk level for human exposure to each individual carcinogen; (ii) 10 times the acceptable risk level for human exposure to each individual noncarcinogen; or (iii) 10 times the acceptable risk level for exposure of individual ecological receptors or populations of ecological receptors to each individual hazardous substance. (B) Are reasonably likely to migrate to such an extent that the conditions specified in subsection (a) or paragraphs $(b)(A)$ or (b)(C) would be created; or (C) Are not reliably containable, as determined in the feasibility study.

3.3.1 Groundwater Beneficial Use Determination

A beneficial water use determination (BWUD) was prepared as part of the 2000 remedial investigation and updated as part of the 2018 supplemental remedial investigation. The findings of the 2000 BWUD indicated that beneficial water uses were limited to aquatic habitat, recreation, and aesthetic quality. Surface water could be used to provide irrigation and/or industrial process/cooling water in the future.

No direct beneficial uses of groundwater occurred at or adjacent to the Site. Potable water at the Site is provided by the City of St. Helens' municipal system. According to the Oregon Water Resource Department database, some domestic wells are present within 1 mile of the Site, none of which are located within the LOF. The nearest wells are located approximately 0.5 mile crossgradient from the Site.

The BWUD concluded that based on the poor yield of the shallow aquifer (less than 0.5 gallon per minute) and the availability of municipal water supply, future direct beneficial uses of groundwater within the LOF are not expected.

Regarding the cleanup of sediments in Milton Creek, Scappoose Bay, and the Multnomah Channel, the reasonably likely future beneficial use of groundwater is limited to discharge through riverbank seeps and/or sediments which provides habitat for plants, mammals, birds, fish and/or benthic organisms.

3.3.2 Surface Water Beneficial Use Determination

Surface water in the vicinity of the Site includes Milton Creek, Scappoose Bay, and the Multnomah Channel. Groundwater likely discharges to surface water in the area of the Site. Surface water within the LOF is currently used for recreation and aquatic and terrestrial habitat. The Boise Cascade (ECSI No. 0014) facility downstream of the Site was identified as using surface water for facility processes. Plant operations have significantly decreased since 2015 with only tissue manufacture remaining. The use of surface water in current operations is unknown but would be significantly less than when pulp and paper operations were active. Reasonably likely future beneficial water uses were determined to include recreational, habitat, and limited industrial use.

3.3.3 Hot Spots

This ROD evaluates soil, groundwater, sediment, porewater, and surface water potential hot spots in the Area 1 Upland and the four in-water PAAs.

3.3.3.1 Non-Aqueous Phase Liquid

Creosote NAPL is present in the subsurface soil beneath the former wood treatment operations area and represents a hot spot. Creosote NAPL at the Site has migrated from the historical ground surface in the former wood-treating operations area to deeper soil and groundwater, affecting nearshore sediment, porewater, and surface water adjacent to Area 1 Upland.

The presence of NAPL in monitoring well MW-3A and downgradient riverbank seeps represents a mobile hot spot to the Cove Area. Riverbank seeps exhibiting a moderate to heavy petroleum sheen are considered mobile hot spots that serve as an ongoing source of contamination to sediment, porewater, and surface water in Upper Milton Creek and Cove Area.

3.3.3.2 Impacts to Beneficial Water Uses

Riverbank groundwater seeps exhibiting a visible creosote sheen is an indication of unacceptable direct contact risk to human health and ecological receptors and is considered a mobile hot spot that serves as an ongoing source of contamination to sediment, porewater, and surface water. Dissolved phase creosote constituents in porewater and surface water that cause unacceptable risk to benthos and fish are considered hot spots. Further assessment during remedial design will be needed to determine if total dissolved creosote constituents in groundwater is a hot spot.

3.3.3.3 Soil/Sediment

The inferred extent of creosote NAPL in upland subsurface soil, riverbank seeps, and near-shore surface sediment represent a soil/sediment hot spot. Subsurface sediment exhibiting a moderate to heavy petroleum sheen may be considered a hot spot if it is not reliably containable. Subsurface sediment exhibiting a creosote sheen outside the in-water PAAs boundaries appears to be reliably contained and is not currently considered a sediment hot spot.

3.3.3.4 Contaminated Debris

Creosote contaminated wood debris is present in surface sediment near and beneath the former overwater operations. Moderate to heavy petroleum sheens are easily generated by disturbing the surface sediment containing this wood debris. Therefore, shoreline and offshore areas of surficial creosote contaminated wood debris is considered a sediment hot spot. Subsurface creosote contaminated wood debris appears to be reliably contained and is not considered a sediment hot spot.

3.4 ESTIMATE OF CONTAMINATION VOLUME

Estimated volumes of contaminated media including soil, sediment, and wood debris are presented on Table 6.

This ROD is based on technical documents that have been reviewed by a technical team at DEQ. The team consists of the project manager, a hydrogeologist, environmental engineer, and human health and environmental toxicologist. The team unanimously supports the selected remedial action.

5.1 REMEDIAL ACTION OBJECTIVES & PRELIMINARY REMEDIATION GOALS

The process for selection of a remedial action by DEQ is outlined in OAR 340-122-0090, focusing on selection of an action that: a) is protective of present and future public health, safety, and welfare of human health and the environment; b) is based on balancing of remedy selection factors; and c) satisfies requirements for hot spots of contamination. DEQ's Guidance for Conducting Feasibility Studies (1998, updated 2006 and 2017) provides more detailed guidance on the remedy selection process, including the development of RAOs, identification of general response actions, identification and screening of remedial technologies, and assembly of RAOs for evaluation. Each of these steps were considered by DEQ, and discussed below, in the selection of a remedial action for the Area 1 Upland and four sediment in-water PAAs.

RAOs and acceptable risk levels, as defined in OAR 340-122-0115(1) through (6), were developed based on the identified beneficial uses, exposure pathways, and the findings of the risk assessments. RAOs are media-specific goals for protecting human health, safety, and the environment and were developed to address the standards established in OAR 340-122-0040. Specifically, the remedial action must achieve the numeric standards for protectiveness that correspond to acceptable risk levels; treat or remove hot spots to the extent feasible; prevent or minimize future releases and migration of hazardous substances in the environment; and provide long-term care or management as necessary and appropriate.

RAOs provide the framework for developing and evaluating RAAs, as any remedy DEQ selects or approves must achieve these Site-specific goals. The RAOs are listed in Section 5.1.2.

5.1.1 Acceptable Risk Levels

Risk-based concentrations (RBCs) and screening levels used to draw conclusions about potential risks to human and ecological health are described in detail in the HHRA and ERA reports. Protectiveness levels were developed for different media at the Site to guide remedy selection during the Feasibility Study (GeoEngineers 2022, Feasibility Study, Section 2.8.2). These protectiveness levels were established to ensure humans and ecological receptors were protected from the potential risks identified in the risk assessments that are summarized in Section 3.2.5 above. Specifically, the Feasibility Study established acceptable risk levels for: PCBs and arsenic in upland soil to protect construction workers and industrial workers working in the upland; naphthalene in groundwater to protect industrial and/or occupational workers from vapor intrusion into buildings; carcinogenic PAHs and TPH in sediment to protect subsistence fishers; and PAHs in porewater and surface water to protect ecological receptors. The Feasibility Study also establishes that the presence of NAPL and/or moderate to heavy petroleum sheen present unacceptable risk to humans and ecological receptors in upland soil, groundwater, sediment, porewater, and surface water (GeoEngineers 2022, Feasibility Study, Section 2.8.2). CULs to inform decision-making during and after remedial implementation will be developed during the remedial design process.

5.1.2 Site-Specific Remedial Action Objectives

Site-specific RAOs were developed for soil and groundwater in Area 1 Upland and sediment, porewater, and surface water in the four sediment in-water PAAs. The RAOs describe what the remedial action is expected to achieve to protect human health, ecological receptors, and beneficial uses, as required by OAR 340-122-0040. The RAOs for the Site are as follows:

Area 1 Upland PAA RAOs:

- RAO 3: Prevent recontamination of in-water remedy from adjacent contaminated riverbank soils.
- RAO 4: Prevent recontamination of in-water remedy from the defined NAPL riverbank seep areas and contaminated groundwater discharge.
- RAO 5: Protect construction and/or excavation worker receptors when working in upland area(s) of inferred NAPL and/or near monitoring well boring MW-10.
- RAO 6: Protect industrial workers in areas of unacceptable risk related to PCB contamination found in gravel roadways.
- RAO 7: Prevent direct exposure to groundwater at Area 1 by Site workers.
- RAO 8: Protect Site workers from vapor intrusion into buildings that cause unacceptable risk.

In-Water Sediment PAA RAOs:

- RAO 1a: Prevent releases of creosote NAPL, defined as moderate to heavy petroleum sheen, to aquatic environment.
- RAO 1b: Prevent unacceptable risk to recreational users associated with direct contact with NAPL in riverbank seeps and surface sediment.
- RAO 1c: Prevent unacceptable risk to aquatic receptors associated with direct contact with NAPL in riverbank seeps, surface sediment, and surface water.
- RAO 2: Protect aquatic receptors from exposure to contaminants in surface sediment (as defined as the top 12 inches of sediment and associated porewater) and surface water that result in toxic effects.

5.1.3 Preliminary Remediation Goals

PRGs were developed for Area 1 Upland and the four sediment in-water PAAs using available Site information to identify endpoint concentrations or risk levels that are believed to be sufficiently protective of human health and the environment. The PRGs describe Site conditions or concentrations in specific media that are designed to be protective of certain receptors from exposure to COCs through particular pathways within a reasonable amount of time. The PRGs for the Site are as follows:

Area 1 Upland PRGs

- Protect construction and excavation workers from direct exposure to NAPL and contaminated groundwater.
- Remove, contain, or treat surface (0 to 3 feet bgs) soil with total PCBs greater than 0.74 mg/kg.
- Protect building occupants from vapor intrusion that results in unacceptable risk.

In-Water Sediment PAAs PRGs

- Remove, contain, or treat NAPL, including heavy to moderate sheen, in surface sediment.
- If not contained, remove, contain, or treat NAPL, including heavy to moderate sheen, in subsurface sediment.
- Remove, contain, or treat NAPL, including heavy to moderate sheen, in shoreline seeps.
- Achieve $\Sigma IWTU$ less than (\leq) 1 in surface sediment porewater within 10 years of removal, containment, or treatment of NAPL in surface sediment and riverbank seeps.
- Achieve all applicable water quality criteria in surface water within 10 years of removal, containment, or treatment of NAPL in surface sediment and riverbank seeps.

5.1.4 Remedial Action Levels and Cleanup Levels

RAOs and PRGs were developed to inform technology evaluations for Area 1 Upland and the four Sediment PAAs. Remedial action levels (RALs) are contaminant-specific concentrations used to identify where an active remedial technology should be applied to reduce risks more effectively than MNR alone. CULs are contaminant-specific concentrations that should not be exceeded following remedial action. RALs and CULs will be developed during remedial design, if necessary.

5.1.5 Priority Action Areas

Area 1 Upland and four sediment in-water PAAs were established for the Site (see Figure 2). PAAs were identified as areas that are highly likely to require active remediation and are characterized by:

• NAPL and/or moderate to heavy petroleum sheen in surface sediment and riverbank seeps.

• Contaminant concentrations that result in unacceptable human and ecological risk with a high degree of certainty.

The Area 1 Upland and four sediment in-water PAAs are discussed in more detail in Section 2.1.1.

5.2 REMEDIAL ACTION ALTERNATIVES FOR SEDIMENT

RAAs were developed and evaluated during the Feasibility Study, and presented in DEQ's Staff Report along with the recommended remedial action. The RAAs are based on general response actions that include: 1) no action, 2) engineering controls and/or ICs, 3) treatment, 4) excavation and offsite disposal, 5) excavation and on-site disposal, 6) and any combination of the general response actions, as appropriate. Several remedial action technologies were evaluated for each general response action. The technologies were screened in accordance with OAR 340-122- 0085(4), which requires meeting the threshold criterion of protecting human health and the environment and considering their relative merits/drawbacks with respect to the remedy selection factors. Tables 3, 4, and 5 provide the rationale for eliminating or carrying forward general response actions and technologies based on Site characteristics, environmental media conditions, and contaminant type. Viable response actions and technologies that can meet the threshold criterion were assembled into RAAs. Given the Site contains hot spots of soil and sediment contamination in Area 1 Upland and in-water areas, the Feasibility Study included an evaluation of a treatment based alternative and/or an excavation and offsite disposal alternative per OAR 340- 122-0085.

5.2.1 **General Response Actions and Applicable Technologies**

Technologies that were carried forward after the initial screening and combined to develop comprehensive RAAs are summarized below:

Engineering and/or Institutional Controls. Engineering controls are physical measures to prevent or minimize exposure to hazardous substances or reduce the mobility or migration of hazardous substances. ICs include legal or administrative actions to reduce exposure to hazardous substances. IC examples include land use restrictions, long-term Site management plans (cap inspection/maintenance and contingency plans), and public access restrictions.

Natural Recovery. Natural recovery relies on ongoing, naturally occurring processes such as sedimentation, biodegradation, and dispersion to contain, reduce, or destroy the toxicity or availability of the contaminants. Monitored natural attenuation (MNA) includes monitoring of these natural processes to assess the rate at which the contaminant concentrations are being reduced. MNA does not include active remedial measures. ENR is the process of accelerating MNA, typically through addition of a thin layer of clean material (e.g., sand). However, natural recovery should not be considered as a viable treatment option for hot spots.

Treatment. Treatment is the permanent and substantial elimination or reduction in the toxicity, mobility, or volume of hazardous substances with the use of in-situ or ex-situ remedial technologies. In-situ treatment for contaminated soils may include solidification/stabilization, enhanced bioremediation, or phytoremediation. In-situ treatment for contaminated sediments may include the addition of reactive materials to conventional sediment caps, such as organoclay, activated carbon, biochar, or an oleophobic biobarrier. Ex-situ technologies for contaminated soils or sediments may include solidification/stabilization or thermal treatment.

Containment. Containment of soils and sediments includes capping, an engineering control that involves the placement of material over the contaminated area. Cap material can be tailored to Site-specific needs. Common cap types include engineered, armored, and reactive caps. The primary functions of a cap are: 1) physical isolation of the contaminated sediment from human and ecological receptors; 2) stabilization of contaminated soil or sediment; and 3) reduction of the flux of dissolved contaminants into the water column (sediments). Capped areas can be engineered in a manner to achieve long-term stability, which may require additional reinforcement. Containment of groundwater can include hydraulic containment using pumping systems, impermeable barrier walls, or permeable reactive barriers that removes contaminants as water flows through.

Removal and Disposal. This technology involves the physical removal (full or partial) of contaminated soil and sediment by excavation or dredging. Material may be disposed offsite to a permitted landfill that is authorized for such disposal under state and federal law. Disposing material offsite would prevent mobility and minimize risk to receptors. Material may also be disposed on-site in a pre-determined upland consolidation disposal facility and managed according to local, state, and federal law. For sediments, some form of dewatering is typically required prior to disposal.

5.2.2 Estimate of Contaminated Media

The estimates for the quantities of contaminated media considered for each of the general response actions above is summarized in Table 6.

5.2.2.1 Estimate of Upland Contamination

The upland source area is approximately 3.45 acres (150,370 square feet) and is currently covered by an average depth of 8 feet of clean fill. Below the clean fill, creosote NAPL in present in the subsurface soil between depths below 8 ft bgs and 25 feet bgs, which results in an estimated total volume of 95,000 cubic yards of creosote-impacted soil.

5.2.2.2 Estimate of In-water Contamination

Area 1 Dock PAA: Wood-related creosote NAPL in surface sediment underneath and surrounding the previous dock structures has been defined by an area approximately 1.85 acres (80,586 square feet). The area has an estimated 435 timber piles. The assumption is these piles are an average of 25 feet long, resulting in a total estimate of 163 tons of wood waste. The estimated depth of the area where hot-spot material is present in the shallow surface sediment with the impacted surficial woody debris is approximately 3,000 cubic yards. In general, the concentrations increase with depth in this area, with elevated concentration reaching depths of 11 feet bml in some area, resulting in a total volume of impacted sediment of 32,800 cubic yards of sediment.

Area 2 Dock PAA: The creosote-impacted wood debris in surface sediments in the Area 2 Dock PAA is approximately 0.34 acres (15,000 square feet) and the total volume of the shallow (mudline to 1 foot bml) impacted sediment is approximately 560 cubic yards. The area has an estimated 235 piles. The assumption is these piles are approximately 25 feet in length and result in approximately 88 tons of wood waste. The combined total for the estimated sediment and wood waste is 519 tons of hazardous waste and 222 tons of non-haz (including the wood piles). In general, the elevated sediment concentrations and presence of moderate to heavy sheen coincides with the surficial debris.

Cove Area PAA: The extent of nearshore surface sediment (up to 1 foot bml), which are largely soft and fine-grained, impacted by creosote NAPL in the Cove Area covers an area of approximately 1.2 acres (52,275 square feet) and is estimated to be approximately 1,950 cubic yards. Moderate to heavy sheen has been identified along approximately 300 linear feet of the shoreline, with steep slopes across approximately 500 linear feet of the riverbank. The sediment concentrations decrease with depth into the subsurface, with depths of contamination identified up to 12 feet bml, with an average depth of 5 feet bml resulting in an overall volume of impacted sediment of 17,200 cubic yards of sediment.

Upper Milton Creek PAA: The area of Upper Milton Creek where moderate to heavy sheen has been identified in the near surface (up to 1 foot bml) sediment and adjacent riverbank, extends approximately 200 linear feet and within an area of less than 0.05 acres (2,200 square feet) The estimated volume of soil and sediment that would be removed to regrade the bank along the creek from approximately 2H:1V to 4H:1V is approximately 1,020 cubic yards.

5.2.3 Common Elements and Assumptions

Area 1 Upland and In-Water Common Elements and Assumptions

The technologies assembled into alternatives represent a spectrum of potential remedial strategies, ranging from capping and natural recovery to complete removal options. Common elements and assumptions to the Area 1 Upland and in-water RAAs are included below and may not be specifically included in descriptions of RAAs.

- Institutional controls. ICs include administrative and legal mechanisms, such as EES, CMMPs, land use restrictions, and water use restrictions to reduce risk to human health, ensure long-term protectiveness of cleanup actions. The purpose of institutional controls is to provide notification regarding the presence of COCs, regulate the disturbance and management of contaminated materials, and aid in the long-term care of cleanup action, including long-term monitoring. If RAOs and PRGs are not completely achieved following remedial action, interim ICs may be necessary, such as a fish advisory.
- Engineering controls. Engineering controls include physical measures to prevent or minimize exposure to hazardous substances in contaminated materials that remain on Site. Examples of engineering controls include measures to prevent access, including fencing, paving pervious surfaces, directing stormwater away from contaminated media, and installing vapor barriers beneath future buildings constructed in upland contaminated areas.
- Remedial design investigations. Additional data collection and evaluations inform remedial design, including but not limited to additional chemical contamination characterization, debris and capping evaluations, measurements of groundwater seepage and river currents, hydraulic and erosion modeling, geotechnical investigations, and seismic design considerations. Preparation and implementation of a comprehensive predesign investigation (PDI) inform Remedial Design/ Remedial Action (RD/RA) as well as support post-performance monitoring activities and post-construction residual risk assessment.
- Recontamination potential. RD/RA addresses recontamination potential of a constructed sediment remedy. Potential source control pathways are further examined during the PDI.
- Post-construction, long-term monitoring. Long-term monitoring is conducted following implementation of remediation actions to ensure remedial actions continue to be protective of human health and the environment and perform as designed and constructed in accordance with RAOs. Long-term monitoring is conducted in perpetuity for isolation walls and caps placed on upland soil and riverbanks and in-water areas. Performance monitoring includes but is not limited to visual inspections for petroleum sheens, cap integrity testing, sediment porewater/surface water sampling, and MNR assessment within and outside the PAAs where applicable to demonstrate sufficient rate of degradation is occurring. Long-term monitoring for natural recovery is conducted until all COC concentrations in sediment/riverbank soil, groundwater/porewater, and surface water are less than established CULs. Updates to appropriate management strategies are based on results of long-term monitoring.
- Permitting. Permitting requirements include endangered species act consultation, biological opinion, 404 USACE/Division of State Lands (DSL) permit, National Pollutant Discharge Elimination System (NPDES) permit, DSL lease negotiations, and applicable floodplain development permits.
- Construction considerations. Remedy construction considerations include evaluated vegetation removal, riverbank stabilization, cap construction, construction management and oversight, and water quality monitoring.
- Achievement of RAOs and PRGs. RAOs and PRGs should be achieved at construction completion, or the shortest reasonable period of time.
- In-water work window. In-water work is performed in the allowable Oregon Department of Fish and Wildlife work window of July 1 to October 31 and December 1 to January 31. The Multnomah Channel is tidally influenced.
- Debris removal. Debris is removed, to the extent necessary or practicable, in areas where applicable, and transported and disposed of at an off-site facility. Similarly, remnant wood pilings are typically cut at the mudline where active remedial technologies are applied. Removed or cut pilings are disposed of at an off-site facility.

In-Water Common Elements and Assumptions

In-water PAAs require active remedy consideration. Multiple objectives were considered during the Feasibility Study for development of in-water alternatives, including integrating chemical isolation and physical protection goals during cap design; habitat enhancements; and slope angle effects on remedy implementation and monitoring, stope stability, wave and wake impacts on the shoreline, and habitat creation.

The following are common elements of all RAAs for the in-water PAAs (except the No Action alternatives) and may not be specifically included in the RAA descriptions.

- Land lease and easement for nearshore contaminated sediment capping.
- Site use restrictions (e.g., limit access to riverbank areas, post signage informing presence of contamination, communication outreach).
- Sediment Management Plan.

In addition to elements listed above, the following are elements of all in-water RAAs that include caps and may not be specifically included in the RAA descriptions:

- The design of all caps need to consider the effectiveness of containing contamination from underlying materials and the potential for recontamination from upland sources.
- Various evaluations are needed to support remedial design, including but not limited to chemical transport modeling, sensitivity analyses, chemical isolation evaluations, erosion protection evaluations, habitat enhancement evaluations, debris and capping evaluations, measurements of groundwater seepage and river currents, hydraulic and erosion modeling, flood impacts, sea level rise resilience, geotechnical investigations, and seismic design considerations.
- The remedial design determines the specific materials, blends, thicknesses, extent, armoring stone sizing, and grading needed to achieve RAOs. Remedial design would result in the preparation of cap design plans and specifications.

5.2.4 RAA-1: No Action

All Upland and In-Water PAAs include RAA-1 (No Action). A "no action alternative" is included for comparative purposes only as stipulated in OAR 340-122-0085(2) and DEQ guidance. Under this RAA, no actions to treat, remove, or monitor COCs would be performed. There would be no reduction in site risk, and thus this RAA is not considered protective by DEQ.

5.2.5 Area 1 Upland Priority Action Area Remedial Action Alternatives

The following summarizes the RAAs considered for the Area 1 Upland PAA, with detailed descriptions of the remedial technologies included in Table 7. Creosote NAPL observed in Area 1 Upland riverbanks adjacent to Milton Creek and Scappoose Bay indicates that some lateral movement of NAPL occurs beneath the Site. As a result, much of the RAAs include source control measures.

RAA-2 fulfills RAOs 5 through 8 by protecting Site workers. Though RAA-2 reduces contaminant mass flux, it does not prevent recontamination of the in-water remedy from adjacent contaminated riverbank soils (RAO 3) or riverbank NAPL seeps (RAO 4). RAA-3 through RAA-6 fulfill RAOs 3 through 8 by protecting Site workers and preventing recontamination of the in-water remedy by controlling contaminant mass flux towards Milton Creek and/or the Cove Area of Scappoose Bay.

- RAA-2: Impervious surface cap and MNA (\$5,565,000).
- RAA-3: Impervious surface cap, hydraulic containment, enhanced bioremediation, and MNA (\$15,097,000).
- RAA-4: Impervious surface cap, permeable reactive barrier, and MNA (\$7,890,000).
- RAA-5: Impervious surface cap, impermeable isolation wall, and MNA (\$8,494,000).
- RAA-6: Excavation and offsite disposal of NAPL area, impervious surface cap, and MNA (\$60,386,000).

5.2.6 In-Water Priority Action Areas Remedial Action Alternatives

5.2.6.1 Area 1 Dock Priority Action Area

The following summarizes the RAAs considered for the Area 1 Dock PAA, with detailed descriptions of the remedial technologies included in Table 8.

The Area 1 Dock RAAs address RAOs 1 and 2 by preventing releases of creosote NAPL to the aquatic environment, preventing risk associated with direct contact by recreational users and aquatic receptors, and protecting aquatic receptors from exposure to contaminants in surface sediment and surface water.

- RAA-2: Armored reactive cap (\$5,382,000).
- RAA-3: Nearshore removal action, upland consolidation, and armored reactive capping (\$8,138,000).
- RAA-4: Nearshore removal action, offsite disposal, and armored reactive capping (\$9,528,000).
- RAA-5: Complete removal and offsite disposal (\$29,181,000).

5.2.6.2 Area 2 Dock Priority Action Area

The following summarizes the RAAs considered for the Area 2 Dock PAA, with detailed descriptions of the remedial technologies included in Table 9.

The following Area 2 Dock RAAs address RAOs 1 and 2 by preventing releases of creosote NAPL to the aquatic environment, preventing risk associated with direct contact by recreational users and aquatic receptors, and protecting aquatic receptors from exposure to contaminants in surface sediment and surface water.

- RAA-2: Armored reactive cap (\$1,309,000).
- RAA-3: Nearshore removal action, upland consolidation, and an ENR sand cap (\$1,596,000).
- RAA-4: Nearshore removal action, offsite disposal, and an ENR sand cap (\$1,604,000).

5.2.6.3 Cove Area Priority Action Area

The following summarizes the RAAs considered for the Cove Area PAA, with detailed descriptions of the remedial technologies included in Table 10.

The following Cove Area RAAs address RAOs 1 through 4 by preventing releases of creosote NAPL to the aquatic environment; preventing risk associated with direct contact by recreational users and aquatic receptors; protecting aquatic receptors from exposure to contaminants in surface sediment and surface water; and preventing recontamination of the in-water remedy from adjacent riverbank contaminated soils, NAPL riverbank seeps, and contaminated groundwater discharge.

- RAA-2: Armored reactive cap (\$3,551,000).
- RAA-3: Riverbank restoration, nearshore removal action, upland consolidation, and armored reactive capping (\$4,836,000).
- RAA-4: Riverbank restoration, nearshore removal action, offsite disposal, and armored reactive capping (\$5,984,000).
- RAA-5: Complete removal and offsite disposal (\$17,606,000).

5.2.6.4 Upper Milton Creek Priority Action Area

The following summarizes the RAAs considered for the Upper Milton Creek PAA, with detailed descriptions of the remedial technologies included in Table 11.

The Upper Milton Creek RAAs address RAOs 1 through 4 by preventing releases of creosote NAPL to the aquatic environment; preventing risk associated with direct contact by recreational users and aquatic receptors; protecting aquatic receptors from exposure to contaminants in surface sediment and surface water; and preventing recontamination of the in-water remedy from adjacent riverbank contaminated soils, NAPL riverbank seeps, and contaminated groundwater discharge.

- RAA-2: Armored reactive cap (\$898,000).
- RAA-3: Regrade streambank, limited removal action, upland consolidation, and armored reactive capping (\$1,146,000).

• RAA-4: Regrade streambank, limited removal action, offsite disposal, and armored reactive capping (\$1,216,000).

5.2.7 Monitoring, Review, and Contingency Plan

There are numerous sources of uncertainty at the Site that make it difficult to predict the long-term effectiveness of any of the RAAs described above, including:

- Heterogeneity in the subsurface.
- Potential changes in future groundwater or surface water use patterns (i.e., beneficial uses).
- Potential changes in future land use and zoning.
- Changes in community concerns regarding remedial actions at the Site.
- Long-term performance of active treatment and/or cap areas.
- The long-term potential for deposition, erosion, or net-neutral conditions in the Sediment Area.
- Factors related to climate change (i.e., rainfall and sea-level rise).

Because of these uncertainties, a Performance Monitoring, Review, and Contingency Plan will be developed that will evaluate the performance of the remedy, and any changes that may affect the ability of the remedy to meet the RAOs. The objective of the Monitoring, Review, and Contingency Plan will be to maintain the overall protectiveness of the selected remedy by establishing a series of decision criteria and related response actions for each potential area of uncertainty identified above, and the RAOs identified in Section 5.1.2 of this document. Section 10.1.8 provides a description of potential contingency measures that could be implemented in the event the RAOs are not achieved following remedy implementation.

6.1 EVALUATION CRITERIA

The criteria used to evaluate the RAAs described below are defined in OAR 340-122-0090 and establish a two-step approach to evaluate and select an RAA. The first step evaluates whether an RAA is protective; if not, the RAA is unacceptable and the second step of evaluation is not required. The RAAs considered protective are evaluated and compared with each other using five balancing factors. The five balancing factors are 1) effectiveness in achieving protection, 2) longterm reliability, 3) implementability, 4) implementation risk, and 5) reasonableness of cost.

An evaluation of how each alternative achieves the preference for treatment or removal of hot spots is also included. Lastly, consideration is given to how each alternative achieves green remediation, as described in DEQ's Green Remediation Policy. The alternative that compares most favorably against the balancing factors and complies with the hot spot criteria is selected for implementation. A residual risk assessment is then conducted for the selected alternative to document that it is protective of human health and the environment.

6.2 PROTECTIVENESS

The protectiveness of a given remedial action is evaluated by assessing whether an alternative would eliminate, reduce, or control risks to human health and the environment and achieve applicable RAOs and PRGs.

OAR 340-122-0090 states that protectiveness may be achieved by any of the following methods:

- Treatment.
- Excavation and offsite disposal.
- Engineering controls.
- \bullet ICs.
- Any other method of protection.
- A combination of the above.

Except for hot spots, there is no preference for any one of the above methods for achieving protectiveness. Where a hot spot has been identified, OAR 340-122-0090(4) establishes a preference for treatment or removal to the extent feasible, including a higher threshold for evaluating the reasonableness of costs for treatment.

Each RAA was screened for whether it is or is not protective. Alternatives that were deemed not to be protective were not evaluated further. Alternatives deemed to be protective were evaluated for the remaining criteria.

6.3 BALANCING FACTORS

The RAAs determined to be protective are evaluated against the following balancing factors defined in OAR 340-122-0090(3) unless otherwise noted:

Effectiveness. Effectiveness is the ability of an alternative to achieve protectiveness. Criteria for evaluating effectiveness include:

- Magnitude of the residual risk from untreated waste or treatment residuals (residual risks), without considering risk reduction achieved through on-site management of exposure pathways (i.e., engineering controls and ICs); the characteristics of the residuals are considered to the degree that they remain hazardous, considering their volume, toxicity, mobility, propensity to bioaccumulate, and propensity to degrade.
- Adequacy of any engineering controls and ICs necessary to manage residual risks.
- The extent to which the remedial action restores or protects existing or reasonably likely future beneficial uses of water.
- Adequacy of treatment technologies in meeting treatment objectives.
- The time until RAOs are achieved.

Each RAA was evaluated and assigned a ranking between 0 and 5 for effectiveness in protecting human health and ecological receptors from contamination risk. A ranking of 0 indicates that an alternative is not effective and a ranking of 5 indicates an alternative provides a high degree of effectiveness.

Long-term reliability. This factor includes evaluation of the ability of an alternative to achieve RAOs over the long-term following remedy implementation. Criteria for evaluating effectiveness include:

- The reliability of treatment technologies in meeting treatment objectives.
- The reliability of engineering controls and ICs needed to manage residual risks, taking into consideration the characteristics of the hazardous substances being managed, the ability to prevent migration and manage risk, and the effectiveness and enforceability over time of the controls.
- The nature and degree of uncertainties associated with any necessary long-term management (e.g., operations, maintenance, monitoring).

Each RAA was evaluated and assigned a ranking between 0 and 5 for long-term reliability in protecting human health and ecological receptors from contamination risk. A ranking of 0 indicates that an alternative has little to no long-term reliability and a ranking of 5 indicates an alternative has a high degree of long-term reliability.

Implementability. This factor includes evaluation of the ease or difficulty in implementing an alternative. Criteria for evaluating implementability include:

- Practical, technical, and legal difficulties and unknowns associated with the construction and implementation of the technologies, engineering controls, and/or ICs, including the potential for scheduling delays.
- The ability to monitor the effectiveness of the remedy.
- Consistency with regulatory requirements, activities needed to coordinate with and obtain necessary approvals and permits from other governmental bodies.
- Availability of necessary services, materials, equipment, and specialists, including the availability of adequate treatment and disposal services.

Each RAA was evaluated and assigned a ranking between 0 and 5 for implementability. A ranking of 0 indicates that an alternative has a significant degree of difficulty in implementing the alternative and a ranking of 5 indicates an alternative has little to no difficulty in implementing the alternative.

Implementation Risk. This factor includes evaluation of the potential risk to human health and the environment associated with remedial action implementation. Criteria for evaluating implementation risk include:

- Potential impacts to the community, workers involved in implementing the remedial action, and the environment and the effectiveness and reliability of protective measures to mitigate these impacts.
- Time until the remedial action is complete.

Each RAA was evaluated and assigned a ranking between 0 and 5 for implementation risk. A ranking of 0 indicates that an alternative has significant implementation risk and a ranking of 5 indicates an alternative has little to no implementation risk.

Reasonableness of Cost. This factor evaluates the reasonableness of the costs associated with the remedial action. Each alternative is assessed for reasonableness of cost by considering:

- Capital, operations and maintenance (O&M), and periodic review costs.
- The net present value of the above.
- The degree to which costs associated with the remedial action are proportionate to the benefits to human health and the environment through risk reduction or management.
- For any hot spots identified, the degree to which costs associated with the remedial action are proportionate to the benefits created through restoration or protection of beneficial uses of water.
- The degree of sensitivity and uncertainty of the costs.

In general, the least expensive remedial action is preferred unless the additional cost of a more expensive corrective action is justified by proportionately greater benefits to one or more of the other balancing factors. A higher threshold is used for evaluating the reasonableness of costs for treatment of hot spots than for remediation of non-hot spot areas.

Each RAA was evaluated and assigned a ranking between 0 and 5 for reasonableness of cost. A ranking of 0 indicates that an alternative has the highest cost and a ranking of 5 indicates an alternative has the lowest estimated cost.

Preference to Treat or Remove Hot Spots. As defined by OAR 340-122-0115(b), nongroundwater or surface water hot spots exist if hazardous substances present an unacceptable risk to human health or the environment and if the contamination is sufficiently concentrated, likely to migrate, or are not reliably containable. Under OAR 340-122-0090(4), a preference is given to alternatives that include treatment or excavation of hot spots to the extent feasible. In addition, a higher cost threshold is applied to sites where hot spots are present.

Each RAA was evaluated and assigned a ranking between 0 and 5 for hot spot treatment or removal. A ranking of 0 indicates that an alternative has no preference to treat or removal hot spots and a ranking of 5 indicates a meaningful reduction in hot spot contamination.

Green Remediation. Green remediation includes practices that lessen the overall environmental impact of remedial actions, such as limiting resources required for implementing the remedy, reducing generation of waste, and reducing greenhouse gas emissions. Alternatives were ranked and evaluated for their inclusion of green remediation technologies or methods.

Each RAA was evaluated and assigned a ranking between 0 and 5 for green remediation. A ranking of 0 indicates that an alternative has no preference for green remediation and a ranking of 5 indicates an alternative has meaningful reduction of resources and greenhouse gas emissions.

6.4 EVALUATION OF ALTERNATIVES AGAINST CRITERIA

Each RAA for each PAA was evaluated for protectiveness and balancing criteria. First, alternatives were screened for whether they achieve or do not achieve protectiveness. Alternatives that do not achieve protectiveness were not further evaluated while those that do achieve protectiveness were evaluated for the remaining criteria. Second, scores and discussion are provided for each balancing factor. Third, initial scores are calculated as the sum of the rankings for effectiveness, long-term reliability, implementability, implementation risk, and cost. Lastly, final scores are calculated as the initial score plus the scores for hot spot treatment and green remediation.

The following sections summarize the outcomes of the scoring and RAA selection for each PAA. Please reference Tables 12 through 16 for a more detailed explanation regarding how each alternative was evaluated against each criterion and Table 17 for a scoring summary for all RAAs considered for each PAA. All alternatives except RAA-1 (No Action) are considered to provide some level of protectiveness and were evaluated for all criteria.

7. COMPARATIVE ANALYSIS OF REMEDIAL ACTION ALTERNATIVES

In this section, each of the RAAs determined to be protective are compared with each other for the remedy evaluation criteria identified in Section 6.1. Because RAA-1 (No Action) is not considered protective and is only used as a baseline for evaluating the other RAAs, RAA-1 is not included in the comparative analysis. Sections 7.1 through 7.5 provide discussion summarizing the major conclusions of the comparative analysis and justification for differentiating issues specific to each of the upland and in-water PAAs. Additional details for each of the PAAs is presented in Tables 12 through 16. Table 17 provides an overall summary of the rankings for each of the PAAs.

7.1 AREA 1 UPLAND PRIORITY ACTION AREA

RAA-4 consisting of an impervious surface cap and permeable reactive barrier had the highest overall score (23) for any of the upland RAAs considered. RAA-6 ranked higher than RAA-4 for effectiveness and long-term reliability because RAA-6 includes removal of the hot spot volume of soil and transporting the material off-site to a permitted facility for disposal. However, during implementation, RAA-6 would have more short-term risks, including limitations related to critical resources, such as dump-trucks that may be limited to meet the demand necessary for completing the project on schedule; higher production of greenhouse gasses; and overall risk to the neighbors as a result of the large increase in traffic. RAA-2 would be easy to implement and therefore scored higher than the other RAAs for implementability and implementation risk. RAA-2 would result in the lowest short-term implementation risks. However, the effectiveness is ranked lower for RAA-2 because it would not meet several of the upland RAOs, including RAO 3 and RAO 4 (prevent recontamination of the in-water remedy). All RAAs would require institutional and engineering controls (e.g., Site use restrictions, CMMP, impervious surface cap, stormwater management, inspections, and maintenance) to achieve RAO 5 through RAO 8.

The cost to complete RAA-2 is the lowest of the RAAs. However, the alternative does not reduce the overall contaminant mass. The cost to complete RAA-4 is slightly less than the cost to complete RAA-5 and both would result in similar removal of upland contaminated soil and be effective in meeting all the upland RAOs (RAO 3 through RAO 8). Although RAA-3 would limit residual mass flux from the upland in groundwater, the cost to implement RAA-3 is higher than RAA-4 and RAA-5 due to the on-going monitoring and maintenance that would be required for the hydraulic containment system. The cost of RAA-6 is an order of magnitude higher than the other RAAs considered for the upland PAA and is considered unreasonable, especially when considering the associated short-term implementation risks.

Additional details and explanation are included in Table 12.

7.2 AREA 1 DOCK PRIORITY ACTION AREA

For the Area 1 Dock PAA, RAA-2 consisting of the installation of an armored reactive isolation cap had the highest initial score (16) considering the five primary evaluation criteria scores and the highest final score (21) when the additional criteria of hot spot treatment and green remedial considerations were also evaluated. RAA-2, RAA-3, and RAA-4 include an armored reactive isolation cap, which will reduce contaminant mass flux to the river and provide a barrier to direct contact with underlying impacted residual or source sediment. RAA-2 scored lower for effectiveness and long-term reliability than RAA-3 and RAA-4, which included removal of shallow hot spot sediment and creosote-impacted timber piles and surficial woody debris. However, RAA-2 scored highest for implementability compared to RAA-3 and RAA-4 because the significant volume of old, highly weathered wood debris present in the Area 1 Dock PAA is unlikely to be successfully removed, would likely make surface and subsurface sediment removal very difficult, and also likely result in an overall increase in the short-term risks to receptors related to the management of the impacted sediment and debris. RAA-2 has less short-term implementation risks to receptors and the environment based on a more focused scope of work, compared to RAA-3 and RAA-4.

Data gaps exist regarding contamination depth profiles in this PAA, and preliminary data suggest that removal of surface material may expose deeper, more highly contaminated material, which was tied into the implementability scores. Additional characterization and evaluation during remedial design would be needed to determine whether removing surface material is feasible and protective.

RAA-2 (amended isolation cap only), is the easiest to implement, has the lowest short-term risks to receptors, and has the lowest cost compared to other alternatives evaluated. However, RAA-2 removes the lowest volume of creosote-contaminated media compared to the other RAAs. RAA-5 would remove the largest volume of hot spot material, has the highest cost, and has the lowest overall score of all the RAAs evaluated, due to challenges associated with implementation and increased implementation risk.

If woody debris can be successfully removed, and subsurface sediment and debris could be accessed, up to approximately 80% of hot spot material could be removed through implementation of RAA-3 and RAA-4, while nearly all hot spot material could be removed with RAA-5. This would result in improved effectiveness and long-term reliability, but the increased scale and complexity of the required construction would make implementation challenging, implementation risks high, and have notably greater costs, particularly for RAA-5. Additional characterization during remedial design will provide additional information to consider if some areas of shallow sediment can be removed without decreasing the effectiveness and implementability in the Area 1 Dock PAA.

Additional details and explanation are included in Table 13.

7.3 AREA 2 DOCK PRIORITY ACTION AREA

RAA-4 scored highest overall for the Area 2 Dock PAA and consists of nearshore removal, offsite disposal, and placing a sand cap for ENR and to cover remaining residuals. RAA-4 and RAA-2 (armored reactive isolation cap only) both had the highest initial score (16). However, RAA-4 had the highest final score (22) because RAA-4 scored higher for effectiveness and long-term reliability through removal of the majority of the contaminated, near-shore surface hot spot material (up to approximately 84%). Each of the alternatives is easy to implement, with local resources available. Additional resources would be required to complete RAA-4 because of the additional scope of work required to dewater the sediment and transport the material to a permitted disposal facility, which resulted in a lower score for implementability and an increase in the shortterm risks to receptors related to off-site transport (pollution, truck traffic through residential areas, etc.). RAA-4 effectively meets DEQ's preference to remove hot spot material from the PAA and removes more hot spot material from the PAA compared to RAA-2. RAA-3 would result in the same volume of hot spot media removed. However, the material would remain on-site, which increases the level of uncertainty for the long-term reliability of an on-site landfill.

The cost to complete the preferred alternative, RAA-4 is slightly higher than the cost to implement RAA-2 and RAA-3. However, the RAA-4 includes the removal of hot spot material compared to RAA-2 and the offsite disposal component included for RAA-4 resulted in a higher score and advantage over RAA-3 because RAA-4 would reduce permitting and stakeholder negotiations and eliminate the engineering controls, ICs, and monitoring and maintenance that would be required for on-site consolidations.

Additional details and explanation are included in Table 14.

7.4 COVE AREA PRIORITY ACTION AREA

RAA-4 consists of riverbank restoration, nearshore removal action, offsite disposal, and an armored reactive cap. RAA-2, RAA-3, and RAA-4 scored the same with an overall score of 21 points for the Cove Area PAA. RAA-3 and RAA-4 scored lower than RAA-2 (armored reactive cap only) for implementability and implementation risks due to their increased scopes. However, RAA-3 and RAA-4 include regrading the riverbank to reduce erosional forces imposed on the cap and remove near-shore shallow sediment, where hot spot concentrations are present, resulting in increased effectiveness and long-term reliability compared to RAA-2. RAA-5 would completely remove all surface NAPL source material and provide improved effectiveness and long-term reliability, but the increased scale and complexity of the required construction would make implementation challenging and result in an overall increase in the implementation risks. RAA-3 scored slightly lower than RAA-4 for the long-term reliability because RAA-3 would require longterm monitoring and maintenance and there is additional uncertainty related to maintaining an onsite landfill.

The cost for RAA-2 was the lowest of all the alternatives evaluated. However, RAA-2 also removes the smallest volume of hot spot material than the other RAAs evaluated. Alternatives RAA-3 and RAA-4 were similar in cost. RAA-5 had a notably greater costs than the other alternatives. Although the final score of RAA-4 was tied with RAA-3, RAA-4 would reduce permitting and stakeholder negotiations and eliminate the engineering controls, ICs, and monitoring and maintenance that would be required for on-site consolidation.

Additional details and explanation are included in Table 15.

7.5 UPPER MILTON CREEK PRIORITY ACTION AREA

RAA-4, consisting of streambank regrading, limited removal action, offsite disposal, and an armored reactive cap had the highest initial score (17) considering the five primary evaluation criteria and tied with RAA-3 for the highest final score (21) when the additional criteria of hot spot treatment and green remedial considerations were also evaluated for the Upper Milton Creek PAA. RAA-3 and RAA-4 both include streambank regrading, which provides many advantages over RAA-2, which does not include regrading. The regrading will remove some hot spot material, facilitate placement of the cap material, improve worker safety, and reduce erosional forces imposed on the cap, thereby resulting in increased effectiveness and long-term reliability compared to RAA-2. The long-term reliability was slightly lower for RAA-3 compared to RAA-4 because there are more future unknowns related to placing material within the 100-year flood plain for RAA-3. The implementation risks for RAA-2 and RAA-3 are less than the short-term risks to complete RAA-4 because RAA-4 would require truck traffic outside of the Site to transport wastes to the landfill. Each of the RAAs are feasible and easy to implement with readily available resources.

The cost to complete each of the RAAs in the Upper Milton Creek PAA are similar, with the cost to complete RAA-2 (armored cap only) slightly less. However, RAA-2 does not meet DEQ's preference for hot spot removal. The cost to complete RAA-4 is the highest of all the alternatives. RAA-4 would reduce permitting and stakeholder negotiations and eliminate the engineering controls, ICs, and monitoring and maintenance that would be required for on-site consolidation.

Additional details and explanation are included in Table 16.
8. RECOMMENDED REMEDIAL ACTION ALTERNATIVES IN STAFF REPORT

After a detailed evaluation of the remedial alternatives for each PAA (see Section 6 and 7), the most feasible, protective alternative was identified and presented to the public for review and comment in the form of a Staff Report. The recommended remedial alternatives are:

- **Area 1 Upland PAA (RAA-4):** This alternative includes placement of an impervious surface cap over the entire Area 1 Upland PAA and a permeable reactive barrier at the top of the riverbank, adjacent to the Cove PAA.
- **Area 1 Dock PAA (RAA-2):** This alternative includes removal of piles and creosoteimpacted surficial woody debris and placement of an armored reactive cap.
- **Area 2 Dock PAA (RAA-4):** This alternative includes removal of surficial woody debris and 235 piles; removal of nearshore contaminated sediment to a depth of 1-foot below the sediment surface, where creosote sheen has been observed; and the placement of a thin, 1 foot-thick sand cap for ENR and to cover remaining sediment that may have residual impacts.
- **Cove Area PAA (RAA-4):** This alternative includes excavation and regrading for hot spot sediment and soil removal and armored reactive capping along the riverbank.
- **Upper Milton Creek PAA (RAA-4):** This alternative includes removal of hot spot soil and sediment, regrading the adjacent bank, installing an armored reactive cap across the Upper Milton Creek PAA, and conducting long-term MNR.

The remedial alternatives for each PAA form an integrated, cost-effective approach that removes and contains contaminated media, including hot spots, through a combination of upland groundwater in-situ treatment using sequestration agents (e.g., organoclays and granular activated carbon amendments); physical isolation of the in-water contamination through a combination of contaminated sediment and debris removal, capping, and contaminant burial (where natural recovery is already occurring); and reduction of risks to receptors through immobilization of bio-available contaminants in the in-water areas.

9. PUBLIC NOTICE, PUBLIC COMMENT, AND AGENCY RESPONSE

Pursuant to ORS 465.320 and OAR 340-122-0100, notice of DEQ's recommended remedial action for the Site was published in the Oregon Secretary of State's *The Oregon Bulletin* June and July issues and bi-weekly in *The Oregonian* and *Columbia County Spotlight* newspapers during June and July. The public notice was also posted on DEQ's public notifications web page. The 60-day comment period commenced on June 1, 2023, and ended on July 31, 2023. Public comments received and DEQ's response are included on Table 18.

DEQ also held several virtual informational sessions with various partners, community members, and stakeholders. Information about the Site, contamination, and recommended remedial action was presented followed by time for questions and answers. One of the informational sessions, minus the questions and answers portion, was recorded and is available to the public.

The Staff Report published May 31, 2023, presented DEQ's recommendation in greater detail based on previous work conducted at the Site, including remedial investigations, risk assessments, and feasibility study activities. These supporting documents are available online through DEQ's ECSI database for the Site [\(ECSI 0959\)](https://www.deq.state.or.us/Webdocs/Forms/Output/FPController.ashx?SourceIdType=11&SourceId=959&Screen=Load).

DEQ's selected remedial action is consistent with the recommended remedial action presented in the Staff Report. The selected remedial action for contaminated media in the Area 1 Upland and contaminated sediments located in Upper Milton Creek and Scappoose Bay adjacent to the former Pope and Talbot wood-treating is protective, and reflects the best balance of effectiveness, longterm reliability, implementability, implementation risks, reasonableness of costs, preference to treat or remove hot spots, and green remediation. Long-term monitoring and maintenance will be required to ensure the remedy remains protective over time. The selected action therefore satisfies the requirements of ORS 465.314 and OAR 340-122-0090.

10.1 DESCRIPTION OF THE SELECTED REMEDIAL ACTION

The following paragraphs describe the selected remedial action for each of the PAAs at the Site. There are no significant changes from the remedial action recommended in the Staff Report and the selected remedial action presented in this ROD. In addition to the specific remedial action described for each PAA below, Site-wide ICs will ensure long-term effectiveness of the selected remedial action for the Area 1 Upland PAA, and in-water ICs and MNR will help ensure effectiveness of the selected remedial action for all of the in-water PAAs. The total estimated cost for the selected remedial action is 22.1 million dollars.

10.1.1 Area 1 Upland Priority Action Area

The selected remedial action for Area 1 Upland PAA is RAA-4, which includes placement of an impervious surface cap over the entire Area 1 Upland PAA and a permeable reactive barrier at the top of the riverbank, adjacent to the Cove area, as conceptually shown in Figure 43. When combined with the Site-wide ICs, the selected remedial action protects on-site workers from direct exposure to the impacted media in the upland and achieves RAOs 5 through 8. The selected remedial action provides source control for the upland groundwater and in combination with the Upper Milton Creek and Cove Area RAAs, also achieves prevention of recontamination of inwater remedies, which will achieve RAOs 3 and 4.

The surface cap would be graded to direct stormwater away from the cap and into a stormwater collection and conveyance system that discharges outside of the Area 1 Upland PAA. Routing stormwater away from the Area 1 Upland PAA reduces infiltration into the underlying subsurface soil, where creosote NAPL is present, and associated groundwater flux towards in-water locations. The impermeable cap also provides protection against direct contact risks to receptors due to PCBs in the roadway soil and provides a vapor barrier to mitigate inhalation risks associated with volatilization of contaminants into ambient air. The PRB will be approximately 2 to 4 feet wide, consist of adsorptive and reactive materials (e.g., organoclay or granular activated carbon), and be placed at the top of the riverbank, upland of the Cove Area, where creosote NAPL seeps have been observed. PRBs have been shown to be effective in treating similar contaminants at other woodtreating cleanup sites. The PRB treatment wall will intercept and sequester creosote NAPL and dissolved phase COCs in groundwater before discharging into the Cove Area of Scappoose Bay, thereby providing source control for the groundwater and potentially mobile NAPL present in the Area 1 Upland PAA. Although hot spots will remain within the Area 1 Upland PAA, the PRB combined with the engineered impervious surface cap and stormwater management system is expected to significantly reduce hot spots mobilizing downgradient of the PRB and stimulate biodegradation of dissolved phase PAHs and TPH in groundwater that migrate towards the Cove Area PAA. The actual footprint of the PRB, type of amendment, and material thickness and depth will be determined during remedial design.

The materials removed during PRB installation will be transported off-site for disposal. The selected remedial action will result in removal of approximately 3,000 tons of hot spot material that will be transported to a subtitle C landfill, and additional approximate 520 tons of impacted material that will be transported for disposal at a subtitle D landfill.

10.1.2 Area 1 Dock Priority Action Area

The selected remedial action for Area 1 Dock PAA is RAA-2, which includes removal of piles and creosote-impacted surficial woody debris and placement of an armored reactive cap. The armored reactive cap will remove direct contact risks to receptors, as identified in RAOs 1a, 1b, and 1c and immobilize bioavailable contaminants in porewater and surface water, which protects aquatic receptors from future exposures of any contaminants remaining below the cap (RAO 2). The Area 1 Dock PAA is shown on Figure 44 and the cross-sections for the selected remedial action are shown on Figures 45A and 45B.

An estimated total of 435 piles will be removed or attempted to be removed from the Area 1 Dock PAA. The piles and surficial debris (approx. 200 tons) will be removed and transported off-site for disposal as non-hazardous waste at a subtitle D landfill. The impacted sediment and debris removed will be transported off-site for disposal. The selected remedial action will result in removal of approximately 160 tons of impacted material that will be transported for disposal at a subtitle D landfill.

Additional sampling will be conducted during RD to inform the actual extent of the cap. Additional sampling will be completed following removal of the debris to verify the leave surface is consistent with the assumptions used for the final amended cap design. DEO anticipates sampling will also determine if some removal of highly concentrated sediment proximal to and underlying creosoteimpacted debris material and in shallow surface sediments is warranted. If found to be necessary, some hot spot surface sediment will be removed for off-site disposal at a permitted landfill, prior to installing the armored reactive cap.

Following removal of the piles and surficial debris, an estimated 1-foot-thick amended isolation cap will be placed across 80,600 square feet (approximately 56% of PAA) of the Area 1 Dock PAA where heavy to moderate sheen is present. The final amendments used in the amended isolation cap will be determined during remedial design. Areas around any remaining piles or dock structures that cannot be removed may require a different formulation of the amendment (e.g., proprietary product), dosage of amendment, or application method to achieve the performance goals of the amended isolation cap. The amended isolation cap will be armored with an estimated 2-foot-thick layer of protective rock with a habitat rock cover.

Surface and subsurface sediment exceeding the hot spot, highly concentrated criteria, will remain in-place, extending in some areas to depths greater than 14 feet bml, unless during remedial design it is determined that removal of this material is feasible and protective. However, this sediment would be isolated from direct contact through installation of the amended isolation cap, and the protective armor cover. Removal of the surficial debris, creosote-impacted timber piles, and installation of the amended isolation cap satisfies the preference for treatment of hot spots to the extent practicable. MNR will be implemented for the Area 1 Dock PAA outside of the limits of the protective isolation cap, with an overall timeline of 10 years to reach the remedial goals for the PAA. If the remedial goals are not met in this time, additional action may be needed.

10.1.3 Area 2 Dock Priority Action Area

The selected remedial action for Area 2 Dock PAA is RAA-4. RAA-4 addresses the in-water RAOs (RAO 1a, 1b, 1c, and 2) and is protective of human health and ecological receptors through a combination of removing surficial woody debris and 235 piles; removing nearshore contaminated sediment to a depth of 1-foot below the sediment surface, where creosote sheen has been observed; and the placement of a thin, 1-foot-thick sand cap over the entire PAA for ENR and to cover remaining sediment that may have residual impacts. The specific cap material, thickness, extent, and sufficiency or insufficiency of using sand only will be determined during remedial design. The Area 2 Dock PAA is shown on Figure 44 and the cross-section for the selected remedial action is shown on Figure 46.

Robust best management practices (BMPs) will be used during the removal of woody debris, piling, and sediment, including the use of a cofferdam and dewatering the area to allow for landbased access to the debris and impacted surface sediment. This management approach during implementation will reduce the short-term risks to in-water receptors and allow for a thorough removal of the sheen-impacted sediment in the Area 2 Dock PAA. The impacted sediment and debris will be transported off-site for disposal. The selected remedial action will result in removal of approximately 500 tons of hot spot material that will be transported to a subtitle C landfill, and additional approximate 200 tons of impacted material that will be transported for disposal at a subtitle D landfill.

10.1.4 Cove Area Priority Action Area

The selected remedial action for the Cove Area PAA is RAA-4, which is a combination excavation and regrading for hot spot sediment and soil removal and armored reactive capping along the riverbank. The amended isolation cap will immobilize bioavailable contaminants in porewater and surface water, which protects aquatic receptors from future exposures of any contaminants remaining below the cap (RAO 2). A depiction of the Cove Area PAA is shown on Figure 44 and the cross-sections for the selected remedial action are shown on Figures 47A, 47B, and 47C.

An estimated total of 139 piles and the hot spot areas of sediment, where heavy sheen has been observed in the Cove Area, will be removed. Removal of the heavy sheens in this area satisfies DEQs hot spot rule and RAOs 1a, 1b, and 1c. Approximately 540 linear feet along the adjacent bank will be regraded to achieve a stable slope of approximately 5H:1V. Removal of the adjacent bank soil will remove additional hot spots and prevent recontamination of in-water remedy from

the adjacent bank area, which satisfies Upland RAO 3. The impacted sediment, bank soil, and debris removed will be transported off-site for disposal. The selected remedial action will result in removal of approximately 3,800 tons of hot spot material that will be transported to a subtitle C landfill, and additional approximate 1,600 tons of impacted material that will be transported for disposal at a subtitle D landfill.

Following removal of the soil and sediment, an estimated 1-foot-thick amended isolation cap will be placed across the Cove Area PAA, estimated to be an area of 50,000 square feet (approximately 33% of PAA) to address the residual soil and sediment concentrations in the areas where the highest concentrations were detected. The final amendments will be determined as part of the remedial design. The amended isolation cap will also be protected by the addition of an estimated 2-foot-thick layer of armor stone with a habitat rock cover. MNR will be implemented for the Cove Area PAA outside of the limits of the cap.

10.1.5 Upper Milton Creek Priority Action Area

The selected remedial action for Upper Milton Creek PAA is RAA-4. RAA-4 includes removal of hot spot soil and sediment, regrading the adjacent bank, installing an amended isolation cap across the Upper Milton Creek PAA, and conducting long-term MNR. The amended isolation cap will immobilize bioavailable contaminants in porewater and surface water, which protects aquatic receptors from future exposures of any contaminants remaining below the cap (RAO 2). Removal of the adjacent bank soil will prevent recontamination of in-water remedy from the adjacent bank area, which satisfies Upland RAO 3. The selected remedial action in Upper Milton Creek PAA also achieves prevention of direct contact with sheen (RAOs 1, 2, and 3). A depiction of the Upper Milton Creek PAA is shown on Figure 44 and the cross-section for the selected remedial action is shown on Figure 48.

Hot spot sediment, including locations where heavy sheen has been observed in the creek, will be removed using land-based equipment. Removal of the heavy sheens in this area satisfies DEQs hot spot rule and RAOs 1a, 1b, and 1c. The streambank adjacent to the creek will be regraded to reduce the current slope (up to 2H:1V) to a more stabilize slope (i.e., 4H:1V). The impacted sediment and impacted bank soil will be analyzed to determine the final disposition of the material. However, it is assumed that up to 1,550 tons of hot spot material will be transported off-site for disposal at a subtitle C landfill and an additional 82 cubic yards of impacted material will be transported for disposal at a subtitle D landfill.

Following removal of the impacted bank soil and creek sediment, an estimated 1-foot-thick isolation cap amended with organoclay will be placed across an approximate 2,000-square-foot area to cover residual soil and sediment. The amended isolation cap will also be protected by the addition of an estimated 2-foot-thick layer of armor stone with a habitat rock cover.

10.1.6 Engineering and Institutional Controls

Until such time that RAOs are achieved, limiting the potential for humans to ingest fish contaminated with primary COCs under recreational angling and subsistence angling scenarios, to the extent practicable, will be necessary. An advisory regarding consumption of fish, shellfish, and crayfish in Multnomah Channel and Scappoose Bay was issued by OHA in December 2020.

Signage communicating the OHA advisory will be installed to dissuade trespassers, recreational anglers, and subsistence anglers from fishing at or near the Site.

Additional ICs will be implemented as needed to manage human health and ecological risks. Other expected ICs that will be considered and more fully addressed during remedial design for the Area 1 Upland PAA include a prohibition on groundwater pumping/use in the upland and a restriction on upland property excavation, new construction, or redevelopment without DEQ approval/authorization. ICs will also be needed for the in-water PAAs, such as restrictions related to dredging within the project limits, vessel anchoring, and limits on navigation over in-water capped areas, etc.

10.1.7 Performance Monitoring

The remedial design work will also be used to inform the requirements for the performance monitoring for the Site, including a determination of points of compliance and timelines for performance monitoring, and an identification of contingency measures that may be implemented in the event that remedial measures prove to be ineffective or do not meet RAOs within specified time frames.

Short-term performance monitoring for the remedial action is anticipated to be completed directly after remedy implementation during Years 0 to 4, and long-term performance monitoring would be performed as determined to be necessary at routine intervals (e.g., Year 5, 10, and 15, etc.) thereafter.

Short- and long-term monitoring activities anticipated for the Site include upland groundwater sampling, in-water multi-beam bathymetric surveys, collection and analysis of surface sediment and porewater sampling, and potentially bioassays and benthic surveys to confirm RAOs are being achieved. Typical performance monitoring programs include a combination of the following monitoring activities:

- Periodic groundwater monitoring to evaluate the concentrations in the vicinity of the Upland PRB are consistent with anticipated design concentrations and the wall is performing as intended (i.e., no excessive biofouling, no excessive movement of the creosote NAPL body is occurring in the vicinity of the PRB, no evidence of active seeps along bank).
- Periodic cap inspections in the upland, along with in-water bathymetric surveys to ensure that the physical integrity of all caps is maintained and not compromised by scour, erosion, or other physical disturbances (e.g., prop-wash, wave-wake effect, etc.).
- Visual observations along Upper Milton Creek and Cove Area PAA banks to ensure the riverbanks remain stable and are not posing a recontamination issue due to excessive erosion and/or chemical breakthrough.
- Sediment, porewater, and/or surface water samples to monitor the overall effectiveness of the in-water remedy and performance of the amended isolation caps over time to ensure that cleanup goals are being achieved within the cleanup time frame (10 years) and

maintained in sediment and porewater at the Site (and if they are not, evaluate whether the Site is being recontaminated by on-site and/or off-site sources).

- Monitor the effectiveness of natural recovery in MNR and ENR areas to ensure that recovery is progressing in areas outside of the in-water PAAs (prioritizing areas where heavy to moderate sheen have been observed) and recovery is progressing in the time frame (10 years) as included in the designed remedial action and identified in the overall PRGs for the Site. This effort should also include further evaluation of areas outside of the inwater PAAs exhibiting light petroleum sheen, especially areas where elevated contaminant levels have been previously observed (e.g., Lower Milton Creek near the location of the 2017 PWS-090617-2 sediment porewater sample).
- Collect data in remedial areas to evaluate contaminant concentration trends in each of the media (groundwater, sediment, and porewater) relative to the final RAOs and PRGs established for the Site during the remedial design effort.

10.1.8 Contingency Measures

In the event that the selected remedial action for each individual PAA is implemented and determined not to meet RAOs, either in the form of upland groundwater sequestration, debris and/or sediment removal, amended isolation capping, or natural recovery, contingency measures may be necessary. If contingency measures are necessary, they are expected to rely on the augmentation of selected remedial action.

Potential contingency measures could include: a) upgrading MNR areas to ENR (i.e., applying sand cover material to MNR and/or ENR areas); b) upgrading ENR areas to in situ treatment (i.e., applying more robust ENR effort in areas by extending the footprint of the ENR area or adding an amendment to the cover); c) upgrading ENR areas to engineered cap; d) increasing cap thickness and/or amendment concentrations; and e) localized hot spot removal with or without capping. Contingency measures will be formally addressed in a Performance Monitoring, Review, and Contingency Plan that will be prepared as part of the forthcoming remedial design document for the property following issuance of this ROD.

10.2 RESIDUAL RISK ASSESSMENT

OAR 340-122-0084(4)(c) requires a residual risk evaluation of the selected remedy to demonstrate that the standards specified in OAR 340-122-0040 will be met, namely:

- Assure protection of present and future public health, safety, and welfare, and the environment.
- Achieve CULs (i.e., the highest of acceptable RBCs or background concentrations).
- For designated hot spots of contamination, evaluate whether treatment or removal is reasonably likely to restore or protect a beneficial use within a reasonable time.

• Prevent or minimize future releases and migration of hazardous substances in the environment.

In the upland, the selected remedy achieves acceptable risk levels through an impervious cap that will limit direct contact with soil. The impervious cap will also limit stormwater infiltration into the upland soil and groundwater source area where hot spot levels of contamination are present, including creosote NAPL. The selected upland remedy also includes a PRB that will limit contaminant migration from the upland source area to the in-water area. The upland remedy does not directly treat source material in upland groundwater or subsurface soil; risks to potential industrial workers, excavation workers, and construction workers in the upland will therefore need to be addressed through ICs (EES and CMMP).

The long-term effectiveness of the selected remedy for Area 1 Upland is dependent on ICs and engineering controls to mitigate risks to potential industrial workers, construction workers, and excavation workers in the upland. The selected remedy for the upland assumes that an EES would be made between the Port of Columbia County and DEQ. ICs and engineering controls, including a CMMP, land use restrictions, water use restrictions, identification of engineered barriers, and long-term monitoring requirements will all need to be incorporated into the EES. To be protective of potential future industrial workers, this EES will need to mitigate or prevent inhalation exposure to contaminants volatilizing to both outdoor and indoor air from groundwater contamination in the source area of the upland. To be protective of potential future excavation workers, this EES will need to mitigate or prevent direct contact with and inhalation of contaminated groundwater in the upland. To be protective of potential future construction and excavation workers, this EES would also need to mitigate or prevent direct contact, ingestion, and inhalation of contaminated subsurface soil in the upland.

In the in-water areas, the selected remedy achieves acceptable risk levels through a combination of removal and capping of contaminated sediment. Concentrations of COCs in surface sediment will be immediately reduced at construction completion. In addition, cleaner sediments from upstream will continue to be deposited, further reducing COC concentrations at the sediment surface. The selected remedy does not, however, result in the immediate removal or destruction of all Site-related NAPL, petroleum sheen, or contaminated woody debris in the in-water areas. Monitoring after remedial implementation will be needed to evaluate any residual risk to subsistence fishers, sport fishers, or ecological receptors associated with residual contamination in sediment, porewater, or surface water in the in-water areas. As described in Section 10.1.8, contingency measures to address any residual risks that are identified will be addressed in a Performance Monitoring, Review and Contingency Plan that will be prepared as part of the forthcoming remedial design document for the Site.

The selected in-water remedy also achieves cleanup risk levels through preventing ongoing releases and direct contact with the NAPL and petroleum sheen at the seep locations and through immobilization of bioavailable contaminants in pore water and surface water in the in-water areas. Although the selected remedy does not result in the actual removal or destruction of all Site-related contaminants, the addition of reactive material to capping material in in-water PAAs is intended to make bioavailable COCs in sediment less bioavailable to humans and ecological receptors. Specifics of cap composition will be identified during remedial design with the intention of reducing bioavailability and mobility of contamination beneath the in-water caps. DEQ anticipates

that porewater data collected post-construction will be used as a LOE to evaluate the efficacy of the remedy in reducing the bioavailability of COCs in impacted in-water areas.

Some potential residual Site risks, primarily associated with light petroleum sheen, are expected for a relatively short period of time over a limited area at the completion of the remedial action. However, the limited residual Site risks are expected to be mitigated through natural recovery processes, and by Year 10, Site-related risks are expected to be within acceptable levels throughout the in-water area. Performance monitoring will track the progress of the preferred remedy in achieving the RAOs following remedy construction in the PAAs. Additional investigation may be needed to evaluate and mitigate any potential residual risks (e.g., associated with light petroleum sheen) in in-water areas outside of the PAAs (e.g., Lower Milton Creek). As discussed in Sections 5.2.7, 10.1.7, and 10.1.8, the performance, monitoring, review, and contingency plans will be developed during remedial design.

The selected remedial action for contaminated media in the Area 1 Upland and contaminated sediments located in Upper Milton Creek and Scappoose Bay adjacent to the former Pope and Talbot wood-treating is protective, and reflects the best balance of tradeoffs considering effectiveness, long-term reliability, implementability, implementation risks, and reasonableness of costs. Long-term monitoring and maintenance will be performed to ensure the remedy remains protective over time. The selected remedial action therefore satisfies the requirements of ORS 465.314 and OAR 340-122-0090.

 $\overbrace{\hspace{2.5cm}}^{+}$

Kevin Parrett, Manager Northwest Region Cleanup Program Department of Environmental Quality

September 1, 2023 Date

ADMINISTRATIVE RECORD INDEX Former Pope & Talbot Wood-Treating Site St. Helens, Oregon

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The Administrative Record consists of the documents on which the selected remedial action for the Site is based. The primary documents used in developing and evaluating RAAs for the former Pope & Talbot Wood-Treating Site are listed below. Additional background and supporting information can be found in the former Pope & Talbot Wood-Treating Site (ECSI No. 0959) project file located at DEQ Northwest Region Office, 700 NE Multnomah Street, Suite 600, Portland, Oregon and [online.](https://ormswd2.synergydcs.com/HPRMWebDrawer/Record?q=webdrawercode%3a%22%2a082%2a%22+And+recAnyWord%3a%22ECSI0959%22&sortBy=recTypedTitle-&pageSize=25&start=1)

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Table 1 Summary of RI Data Collection Updated Supplemental RI Report Former Pope & Talbot Wood-Treating Site St. Helens, Oregon

Sample Count

Please refer to notes at end of table.

Table 1 Summary of RI Data Collection Updated Supplemental RI Report Former Pope & Talbot Wood-Treating Site St. Helens, Oregon

Analytical Tests by Matrix Count

Acronyms/ Abbreviations

PAHs = polycyclic aromatic hydrocarbons

SVOCs = semivolatile organic compounds

voes = volatile organic compounds

NAPL = nonaqueous-phase liquid

RI = Remedial Investigation

Other= Total Organic Carbon, Dissolved Organic Carbon, grain size, and/or percent solids

Notes

Sample Count = number of samples collected from each medium

Analytical Test Count= number of tests run for each chemical class

Analytical Results Count = number of individual analyte results for each medium

Analytical Tests by Matrix Count = number of tests run for each chemical class and each medium

SVOCs include PAHs by various methods; refer to the 2000 RI Report (GeoEngineers 2000) for a full list of SVOC analytes reported in the RI.

Table **2** Summary of Risks Former Pope & Talbot Wood Treating Site St. Helens, Oregon

^A The cancer risk and hazard index for surface water include all samples from Scappoose Bay and Milton Creek; the data were not segregated into - Fish tissue / consumption (sport fisher) - The bioaccumulative PAHs are not carcinogenic.

NME - resonandle manienum exposure

CIE - central tendency exposure

HHRA - human health in's assessment

HHRA - human health ti RME - reasonable maximum exposure

THER - human health risk assessment

The cancer risk and hazard index for surface water include all samples from Scappoose Bay and Milton Creek; the data were n

² The cancer risk and h

Upland and Riverbank Surface Soil Technology Screening Port of Columbia County - Former Pope & Talbot Wood Treating Site St. Helens, Oregon

Table 3

Table 3

Table 3

Note:

Shading indicates technologies that have been eliminated from consideration.

Upland Subsurface NAPL, Soil, and Groundwater Technology Screening Port of Columbia County - Former Pope & Talbot Wood Treating Site St. Helens, Oregon

GEOENGINEERS

Note:

Shading indicates technologies that have been eliminated from consideration.

In-Water Sediment and Porewater Technology Screening Port of Columbia County - Former Pope & Talbot Wood Treating Site St. Helens, Oregon

Note:

Shading indicates technologies that have been eliminated from consideration.

Notes:

¹ See Section 8 of Staff Report for description of preferred RAAs.

 2 Estimates assume 1.6 tons/CY for impacted soil and sediment.

1 See Staff Report Section 5 bgs below ground surface cm/s centimeters per second COCs Constituents of Concern CY Cubic Yards IC **Institutional Control** MNA Monitored Natural Attenuation NAPL **Non-Aqueous Phase Liquid** PAA Priority Action Area PAHs Polycyclic Aromatic Hydrocarbons PRB Permeable Reactive Barrier RAA Remedial Action Alternative by below ground

contineters per constituents c

COCs

CV C C Unic Visual C

C Unstitutional C

Incorded National C

MNA

Mon-Aqueous

PAA

PAHs

Priority Action

PAHs

Priority Action

PAHs

RAA

REA

SF

Square Feet

Ser percent Selected Remedial Action

1 See Staff Report Section 5 bml

CY Cubic Yards

Cubic Yards Cubic Yards PAA Priority Action Area RAA Remedial Action Alternative
SF Square Feet SF
% Square Feet
percent percent **Selected Remedial Action**

Table 9. Area 2 Dock PAA remedial technologies descriptions and assemblage of RAAs.

Table 10. Cove Area PAA remedial technologies descriptions and assemblage of RAAs.

RAA Remedial Action Alternative

SF Square Feet

Selected Remedial Action

Table 12. Area 1 Upland PAA remedial action alternatives rankings and rationale for evaluation criteria.

Table 12. Area 1 Upland PAA remedial action alternatives rankings and rationale for evaluation criteria.

Notes:

1 Staff Report Sections 6.2 and 6.3 provide evaluation criteria considered for each of the balancing factors. 2 Initial score includes the combined scores from the 5 balancing factors described in OAR 340-122-0090(3)

3 The final score includes additional scores from additional preferred criteria, including hot spot treatment and green remediation.

ARARs **Applicable or Relevant and Appropriate Requirements**

COCs Constituents of Concern

IC Institutional Control
IDW Investigation Derived

IDW Investigation Derived Waste
MNA Monitored Natural Recovery

MNA Monitored Natural Recovery
NAPL Mon-Aqueous Phase Liquid

NAPL Non-Aqueous Phase Liquid

O&M Operations and Maintenance

O&M Operations and Maintenance

PAA Priority Action Area

PAA Priority Action Area
PRB Permeable Reactive

PRB Permeable Reactive Barrier
RAA Remedial Action Alternative RAA Remedial Action Alternative
RAA Remedial Action Objective

RAO Remedial Action Objective
TOC Total Organic Carbon Total Organic Carbon

Table 13. Area 1 Dock PAA remedial action alternatives rankings and rationale for evaluation criteria.

Table 13. Area 1 Dock PAA remedial action alt^e Table 13. Area 1 Dock PAA remedial action alternatives rankings and rationale for evaluation criteria.

Notes:

Table 14. Area 2 Dock PAA remedial action alternatives rankings and rationale for evaluation criteria.

Table 14. Area 2 Dock PAA remedial action alternatives ran

- **BMP Best Management Practice** ENR Enhanced Natural Recovery
- O&M Operations and Maintenance
- PAA Priority Action Area
- RAA Remedial Action Alternative
- RAO Remedial Action Objective
- % Percent

Notes:

1 Staff Report Sections 6.2 and 6.3 provide evaluation criteria considered for each of the balancing factors.

2 Initial score includes the combined scores from the 5 balancing factors described in OAR 340-122-0090(3)

3 The final score includes additional scores from additional preferred criteria, including hot spot treatment and green remediation.

Table 15. Cove Area PAA remedial action alternatives rankir Table 15. Cove Area PAA remedial action alternatives rankings and rationale for evaluation criteria¹.

Notes:

Table 16. Upper Milton Creek PAA remedial action alternativ

2 Initial score includes the combined scores from the 5 balancing factors described in OAR 340-122-0090(3)

3 The final score includes additional scores from additional preferred criteria, including hot spot treatment and green remediation.

O&M Operations and Maintenance

PAA Priority Action Area

RAA Remedial Action Alternative

RAO Remedial Action Objective

Notes:

RAA Remedial Action Alternative
PAA Priority Action Area Priority Action Area

Selected Remedial Action

Table 17. Summary of evaluation criteria scoring for each remedial action alternative for all PAAs.

DRAWN BY: SD CHECKED BY: JKH

DRAWN BY: PM CHECKED BY: JKH

- A@ Existing Monitoring Well
- **1** Direct-Push Boring
- \blacktriangle Seep

- ¹ No Petroleum Sheen or NAPL Observation
- \bullet Petroleum Sheen Slight
- **1** Petroleum Sheen Moderate to Heavy (NAPL)
- . Visible Creosote Sheen Observed in Surface Sediments (2010, 2011, 2012)

LEGEND:

March 2011 Phase 1 TarGOST Boring:

- \bigoplus No Petroleum Sheen or NAPL Signal
- \bigoplus Petroleum Sheen Slight
- **C** Petroleum Sheen Moderate to Heavy

July 2011 Phase 2 Sediment Boring:

July/October 2012 Sampling Event:

Porewater Sample Location

- \Diamond Sheen Degree Unknown
- ◆ No Sheen
- ◆ Petroleum Sheen Slight
- ◆ Petroleum Sheen Moderate to Heavy

Limited Sediment Investigation (LSI) Sample Location

 \Diamond Sheen - Degree Unknown

- \lozenge No Sheen
- \Diamond Petroleum Sheen Slight
- Petroleum Sheen Moderate to Heavy

October 2013 Dock Area Investigation:

- ∇ Hand Auger without NAPL Observation
- **E** Pothole without NAPL Observation
- **DECOM** Pothole with NAPL Observation
- **.** DAI Boring without NAPL Observation
- **.** DAI Boring with NAPL Observation
- Surface Water Cage
- Phase 3 Bulk Sediment and Porewater Sample Location
- ----- Former Site Feature
- Inferred Extent of Upland NAPL
	-
- Approximate Upland Site Boundary (Ordinary High Water 14 ft. NAVD 88)
- **-** Cross Section Location

-
-
-
-
-
-

-
-
-
-
-
-

LEGEND:

 \triangleq Existing Monitoring Well

March 2011 Phase 1 TarGOST Boring:

- ♦ No Petroleum Sheen or NAPL Signal
- \bigoplus Minor Signal (Petroleum Sheen Slight)
- ♦ NAPL Signal

- <> Sheen Degree Unknown
- No Sheen

◆ Petroleum Sheen - Slight

◆ Petroleum Sheen - Moderate to Heavy

July 2011 Phase 2 Sediment Boring:

• No Petroleum Sheen or NAPL Observation

O Petroleum Sheen - Slight

• Petroleum Sheen - Moderate to Heavy (NAPL)

July/October 2012 Sampling Event:

October 2013 Dock Area Investigation:

v Hand Auger without NAPL

□ Pothole without NAPL Observation

Pothole with NAPL Observation

C DAI Boring without NAPL Observation

O DAI Boring with NAPL Observation

Inferred Basalt Contour (4-foot interval) ---(darker shading indicates lower elevation)

Former Site Feature

- **L -** Areas of Significant Creosote Contaminated **L** Wood Debris
- Inferred Extent of Upland NAPL

D Approximate Upland Site Boundary (Ordinary | Approximate Upland Site Boundary | Ordinary | Approximate Site Boundary | Condition Site Boundary | Condition Site Boundary | Condition Site Boundary | Condition Site Boundar

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- A@ Existing Monitoring Well
- **1** Direct-Push Boring
- \triangle Seep
- A RI Sample Location
- **ED** Surface Water Level Measurement Station
- **1** October 2011 Shoreline Observation:
- \triangleq March 2011 Phase 1 TarGOST Boring
- **July 2011 Phase 2 Sediment Boring**

LEGEND:

July/October 2012 Sampling Event:

- ◆ Sediment Porewater Sample Location
- \Diamond Limited Sediment Investigation (LSI) Sample Location

October 2013 Dock Area Investigation:

- ∇ Hand Auger
- **D** Pothole
- **.** DAI Boring

September 2017 Sampling Event:

- \bullet Phase 3 Data Gap Investigation Location
- Former Site Feature
-

- A@ Existing Monitoring Well
- March 2011 Phase 1 TarGOST Boring:
- \bigoplus No Petroleum Sheen or NAPL Signal
- \bigoplus Petroleum Sheen Slight

\bigoplus Petroleum Sheen - Moderate to Heavy

- July 2011 Phase 2 Sediment Boring:
-
- \bigodot Petroleum Sheen Slight
- Petroleum Sheen Moderate to Heavy (NAPL) \blacklozenge No Sheen
- . Visible Creosote Sheen Observed in Surface Sediments (2010, 2011, 2012) And Detroleum Shee

LEGEND:

July/October 2012 Sampling Event:

Porewater_Sample_Location

- ◆ Sheen Degree Unknown
- ◆ No Sheen
- ◆ Petroleum Sheen Slight
- ← Petroleum Sheen Moderate to Heavy
- **G** No Petroleum Sheen or NAPL Observation Limited Sediment Investigation (LSI) Sample Location September 2017 Sampling Event:
	- \Diamond Sheen Degree Unknown
	-
	- \Diamond Petroleum Sheen Slight
	- ♦ Petroleum Sheen Moderate to Heavy

October 2013 Dock Area Investigation:

- ∇ Hand Auger without NAPL Observation
- **D** Pothole without NAPL Observation
- **D** Pothole with NAPL Observation
- **.** DAI Boring without NAPL Observation
- **.** DAI Boring with NAPL Observation

C Phase 3 Data Gap Investigation Location (No Sheen)

----- Former Site Feature

- **C** P Areas of Significant Creosote Contaminated Wood Debris
- Inferred Extent of Upland NAPL
- <u>↑↑</u> Area of Intermittent Groundwater/NAPL Seeps
- **Approximate Offshore Areas**
- Approximate Upland Site Boundary (Ordinary High Water 14 ft. NAVD 88)

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HUMAN HEALTH RISK ASSESSMENT (HHRA) CONCEPTUAL SITE MODEL (CSM)

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Notes: 1. The subsurface conditions shown are based on interpolation between

Legend

- Angular Riprap with Voids Filled by Smooth, Rounded Rock (e.g., Fish Mix)
-

50/50 Mixture of Sand and Organoclay

Ordinary High Water Mark (14 ft.)

Vertical Datum: North American Vertical Datum of 1988 (NAVD88)

subsurface conditions may vary from those shown.

will serve as the official document of record.

widely spaced explorations and should be considered approximate; actual

2. This figure is for informational purposes only. It is intended to assist in the

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Legend

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Vertical Datum: North American Vertical Datum of 1988 (NAVD88)

Vertical Scale in Feet
Vertical Exaggeration = 2X

Legend

- Angular Riprap with Voids Filled by Smooth, Rounded Rock (e.g., Fish Mix)
- 50/50 Mixture of Sand and Organoclay

NAPL Seep \rightarrow

Ordinary High Water Mark (14 ft.)

Vertical Datum: North American Vertical Datum of 1988 (NAVD88)

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NAPL Seep \rightarrow

Ordinary High Water Mark (14 ft.)

Vertical Datum: North American Vertical Datum of 1988 (NAVD88)

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NAPL Seep

Ordinary High Water Mark (14 ft.)

50/50 Mixture of Sand and Organoclay

Vertical Datum: North American Vertical Datum of 1988 (NAVD88)

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